

Reliability evaluation of biomarker reference ranges for mesocosm and field conditions: Cellular innate immunomarkers in Gasterosteus aculeatus

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1	Reliability evaluation of biomarker reference ranges for
2	mesocosm and field conditions: cellular innate immunomarkers in
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Due to their sensitivity to environmental contamination and their link with fish health status, 18 innate immunomarkers are of great interest for environmental risk assessment studies. 19 20 Nevertheless, the lack of knowledge about the effect of confounding factors can lead to data misinterpretation and false diagnostics. So, the determination of reference values was of huge 21 22 interest for the integration of biomarkers in biomonitoring programs. Laboratory 23 immunomarker reference ranges (including cellular mortality, leucocyte distribution, phagocytosis activity, respiratory burst and lysosomal presence) that consider three 24 25 confounding factors (season, sex and body size) were previously developed in three-spined 26 stickleback, Gasterosteus aculeatus, from our husbandry. Usefulness of these reference ranges in biomonitoring programs depends on how they can be transposed to various 27 experimental levels, such as mesocosm (outdoor artificial pond) and field conditions. 28 Immunomarkers were therefore measured every two months over one year in one mesocosm 29 and in one site assumed to uncontaminated (Houdancourt, field). Differences between 30 31 immunomarker seasonal variations in *mesocosm* and *field* fish on one side and *laboratory* fish on the other side were quantified: in some cases, seasonal trends were not significant or did 32 not differ between mesocosm and laboratory conditions, but overall, models developed based 33 34 on data obtained in *laboratory* conditions were poorly predictive of data obtained in mesocosm or field conditions. To propose valuable field reference ranges, mesocosm and field 35 data were integrated in innate immunomarker modelling in order to strengthen the knowledge 36 on the effect of confounding factors. As in laboratory conditions, sex was overall a 37 confounding factor only for necrotic cell percentage and granulocyte-macrophage distribution 38 39 and size was a confounding factor only for cellular mortality, leucocyte distribution and phagocytosis activity. Confounding factors explained a large proportion of immunomarker 40 variability in particular for phagocytosis activity and lysosomal presence. Further research is 41

42	needed to test the <i>field</i> models in a biomonitoring program to compare the sensitivity of
43	immunomarkers to the confounding factors identified in this study and the sensitivity to
44	various levels of pollution.

45 Keywords: biomarker reference ranges; fish conditions; innate immunity;

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47 Highlights:

- 48 1- Similar immunomarker seasonal variations for *laboratory*, *mesocosm* and *field* fish
- 49 2- Importance of fish conditions on reference ranges
- 50 3- Better predictability of phagocytosis activity and granulocyte-macrophage percentage
- 51 4- Development of *field* reference ranges for cellular innate immunomarkers

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53 I. Introduction

Biomarkers reflect the physiological state of an organism and can serve to detect exposures to 54 xenobiotics or their early effects. They have therefore been proposed to determine effects of 55 water quality on aquatic wildlife. Nevertheless, to be used currently in biomonitoring 56 campaigns, some criteria, including natural variability (van der Oost et al., 2003), should be 57 considered in data analysis (Flammarion et al., 2000; van der Oost et al., 2003) to reduce data 58 59 misinterpretation. In fact, many confounding factors, including variations of environmental conditions (temperature, salinity, photoperiod or oxygen levels) and organism's morphology 60 61 and life history traits (sex, age or sexual maturation) could greatly influence biomarker 62 responses (Bowden, 2008; Bowden et al., 2007; Uribe et al., 2011). For example, Shepherd et al. (2012) showed differences between sexes in the involvement of the SOCS genes in the 63 vellow perch immune response. Since the immune system interacts in a bidirectional way 64 with the endocrine system, variations in hormone concentrations due to the reproductive cycle 65 of wild fish may be a major confounding factor for the measurement of immune responses. In 66 67 fact, gonadal sex hormones affect innate immune capabilities by acting directly on hormonal receptors or indirectly through non-hormonal receptor mediated mechanisms (Ansar Ahmed 68 2000). In the same way, sex steroids are known to modulate greatly the immune response of 69 70 fish (Slater and Schreck 1993, Yamaguchi et al. 2001, Cabas et al. 2018). Regarding the effect of fish body size on immune response, little information was available. Even so, the 71 attractiveness of fish to parasites was appears to be higher in larger fish notably due to 72 increased contact with water flow. To stop this phenomenon, the host could increase their 73 immune responses (Lo et al. 1998). All these sources of variability may disturb biomonitoring 74 results, limit the interpretation of variation between sites, and may lead to incorrect site 75 76 classification.

To limit the risk of false positive and false negative results, many authors suggest comparing 77 78 biomonitoring results to robust reference values (Barrick et al., 2018, 2016; Burgeot et al., 2010; Kumari et al., 2006; Mohanty et al., 2014; Sanchez et al., 2008). Unfortunately, most of 79 the developed reference values ignore confounding factor effects by providing either only one 80 mean with a confidence interval or a range of means over a sampling period (Kumari et al., 81 2006; Mohanty et al., 2014; Sanchez et al., 2008). To better determine effects of confounding 82 factors on biomarker responses, two methods were proposed. The first method, suggested by 83 the Convention for the Protection of the Marine Environment of the North-East Atlantic 84 (OSPAR) and by the International Council for the Exploration of the Sea (ICES), consists in 85 collecting organisms on a same reference site during several years to integrate effects of many 86 confounding factors (Barrick et al., 2018, 2016; Burgeot et al., 2010). Nevertheless, this 87 method was only representative of the water quality of the site used and it could be difficult to 88 89 separate specific effects of each confounding factor. Within the second method, the influence of each confounding factor is evaluated separately before being integrated in a larger model of 90 91 biomarker responses (Coulaud et al., 2011; Hanson, 2011; Krell et al., 2011; Maltby et al., 92 2002).

Among all developed biomarkers, due to its sensitivity to environmental contaminations and 93 94 its direct link with individual health status, the immune response is considered as an attractive non-specific marker for environmental biomonitoring (Bols et al., 2001) which could help to 95 better identify risks associated to an ecosystem contamination (Bado-Nilles et al. 2014). 96 Among all the measurable parameters composing innate and acquired components of the 97 immune system, biomarkers related to innate immune functions were highly relevant due to 98 99 their response being non-dependent of previous exposure to foreign antigens (Monserrat et al., 100 2007). In teleosts, the innate mechanism of phagocytosis appears to be the central cellular 101 immune process. Mainly two leucocyte sub-populations, the granulocytes and the

monocytes/macrophages ("granulocyte-macrophage percentage"), are responsible for 102 phagocytosis. This process consists in the adhesion to the leucocyte membrane and the 103 engulfment of the target particle, followed by its destruction by a combined action of aerobic 104 destruction pathway by respiratory burst and anaerobic destruction pathway by lytic enzymes 105 contained in lysosomes (Ellis, 1999; Magnadóttir, 2006). In ecotoxicology, each part of this 106 phagocytosis mechanism has been described as an attractive biomarker of field pollution in 107 fish (Betoulle et al. 2000, Bols et al. 2001, Reynaud and Deschaux 2006). For example, the 108 109 phagocytic efficiency was repressed in rainbow trout (Oncorhynchus mykiss) after one week of exposure to Montreal municipal sewage effluent (Salo et al. 2007). In the same way, 110 phagocytic efficiency and capacity were reduced in rainbow trout coming from industrial 111 waste incineration site (Benchalgo et al. 2014). Modifications of respiratory burst index and 112 lysosomal presence were affected by many pollutants (Arnold et al. 1995, Roméo et al. 2000, 113 114 Ahmad et al. 2004, Santos et al. 2006) and were currently associated to cell death (Holtzman 1989, Zdolsek et al. 1990, Risso-de Faverney et al. 2001, Guicciardi et al. 2004, 115 116 Krumschnabel et al. 2005, Kurz et al. 2008). For example, copper induced both necrotic and 117 apoptotic cell death on rainbow trout hepatocytes by stimulation of ROS production, essentially into mitochondria, which favors destabilization of lysosomal membranes 118 (Krumschnabel et al. 2005). In the same way, an increase of ROS production in Nile tilapia 119 120 (Oreochromis niloticus) splenocytes by endosulfan could induce at term lysosomal destabilization and cellular mortality (Tellez-Bañuelos et al. 2009). Induction of cellular 121 mortality and decrease of phagocytosis activity could contribute to an increased susceptibility 122 to opportunistic infections in animals (Arkoosh et al. 1998, Misumi et al. 2005, Bado-Nilles 123 et al. 2009, Kreutz et al. 2010, Danion et al. 2011, Marchand et al. 2017). Nevertheless, as 124 125 previously explained, use of these immunomarkers in biomonitoring campaigns depends on applicability of reference values. Few studies provide information on their natural variability,which partly explains why they are scarcely used by managers in biomonitoring contexts.

128 Reference ranges of some innate immunomarkers were first modelled in function of season, 129 sex, and body size in the three-spined stickleback (Gasterosteus aculeatus) (Marchand et al. 2019). In the present paper, the previously established laboratory reference ranges are 130 evaluated at various experimental levels, ranging from the laboratory to the field. In this way, 131 132 to evaluate the relevance of immunomarker reference ranges previously developed (Marchand et al. 2019), data obtained in stickleback in other conditions, (outside artificial pond: 133 mesocosm; Houdancourt reference site: field) were compared to the laboratory reference 134 ranges. 135

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137 II. Material and Method

a. Experimental level of sticklebacks for each experimental level reference range

During this study, all mature three-spined sticklebacks used for laboratory and outside
artificial pond experiments were obtained from the INERIS husbandry (Verneuil-en-Halatte,
France). The *laboratory* reference ranges used in the present work had been developed based
on data obtained in 282 fish in a previous experiment, from December 2015 to August 2016
(Marchand *et al.* 2019).

To make the *mesocosm* reference ranges, 112 sticklebacks were maintained in one outside artificial pond containing sediments, natural vegetation, phytoplankton, periphyton, and a macroinvertebrate community which were well-suited to stickleback (Jones and Stafford 2019). Every two months, from February 2016 to December 2016, 15 - 20 fish were sampled (Table I). Fish were anaesthetized using tricaine methanesulfonate (100 mg/L, Sigma) thensacrificed, measured, weighed, and the spleen was removed to measure immune responses.

Field reference ranges were developed using sticklebacks sampled at Houdancourt (60, Oise, France – Table I), an uncontaminated site located immediately upstream of a watercress exploitation, from October 2016 to April 2018. At each sampling date, temperature, dissolved oxygen, pH and conductivity were monitored and 20 adult fish were caught by electrofishing, anaesthetized (tricaine methanesulfonate, 100 mg/L, Sigma), sacrificed, measured, weighed, and the spleen was removed to measure immune responses.

Mesocosm and laboratory sampling were made strictly at the same time, whereas fieldsampling was sometimes a few weeks before or after.

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b. Leucocyte isolation and innate immune biomarker analysis

The spleen filters plasma, traps blood-borne substances, and enriches blood with new immune 160 cells (Press and Evensen, 1999), and was therefore selected to assess immunological 161 characteristics. A splenic leucocyte suspension was obtained by gently pressing the spleen 162 through sterilized nylon mesh (40 µm, Sigma) with 5 mL Leibovitz 15 (L15) medium (Sigma) 163 containing heparin lithium (100 mg/L, Sigma), penicillin (500 mg/L, Sigma), and 164 streptomycin (500 mg/L, Sigma). To eliminate any bias due to stressful conditions of the fish 165 before sacrifice, samples were stored during 12 hr at 4°C (Bado-Nilles et al., 2014). Then, a 166 167 Malassez haemocytometer was used to adjust leucocyte concentration in samples to 10^6 cell/mL in L15 medium. Analyses were carried out by flow cytometry, on whole leucocytes, 168 using a CyAnTM ADP (Beckman coulter) flow cytometer. A total of 10,000 events per sample 169 were analyzed after cell excitation by a 488 nm-argon laser. 170

Percentage of each leucocyte sub-population (lymphocyte and granulocyte-macrophage) were
determined by measuring size (forward scatter, FSC) and complexity (size scatter, SSC)
(Bado-Nilles et al., 2014). For ease of reading, only the granulocyte-macrophage percentages
among the total leucocyte populations are presented in figures and tables.

A double labelling, with Yo-PRO®-1 (Invitrogen, final concentration: 3.14 mg/L) and propidium iodure (Invitrogen, final concentration: 5.01 mg/L) probes, was used to assess the cellular mortality percentages. After 10 minutes of incubation on ice and in the dark, cellular fluorescence parameters were measured. Apoptotic and necrotic cells expressed respectively green (FL1) and red (FL3) fluorescences (Bado-Nilles et al., 2014).

180 The measurement of leucocyte respiratory burst was performed using a modification of the Chilmonczyk and Monge (1999) technique, adapted for stickleback. Determination of reactive 181 182 oxygen species (ROS) production in unstimulated cells depends upon the cell incorporating 2',7'-dichlorodihydrofluorescein diacetate acetyl ester (H2DCF-DA, Invitrogen, final 183 184 concentration: 29.30 mg/L), a stable non-fluorescent molecule which was hydrolysed to 185 dichloro-dihydro-fluorescein diacetate (DCFH) by cytosolic enzymes. When leucocytes were stimulated with phorbol 12-myristate 13-acetate (PMA, Invitrogen, final concentration: 186 9.25 mg/L), the most specific inductor of respiratory burst (Ambrozova et al., 2011; 187 188 Chadzinska et al., 2012), H2DCF-DA was hydrolysed by H2O2. Finally, the DCFH obtained was oxidized to the fluorescent dichlorofluorescein (DCF) to enable quantification, after 30 189 190 min of incubation at room temperature, by flow cytometry of unstimulated and stimulated cells in FL1. The respiratory burst index was determined as the ratio of fluorescence of PMA 191 stimulated cells (H₂DCF-DA plus PMA) to that of unstimulated cells (H₂DCF-DA). 192

193 Intracellular lysosomal presence was determined by incubation of spleen leucocyte 194 suspension with 0.3 mg/L of acridine orange (AO, Sigma), a lysosomotropic weak base, for 20 min in the dark at room temperature. Then, fluorescence was measured in FL3 (Bado-Nilles et al., 2013).

Phagocytosis activity was evaluated after 1 h of incubation in the dark and at room temperature of spleen leucocyte suspension with fluorescent microsphere at a concentration of 2.7 x 10^7 particles/mL (Fluorospheres® carboxylate-modified microsphere, diameter 1 µm, Invitrogen) (Gagnaire et al., 2004). Phagocytosis activity was characterized by two biomarkers, phagocytic capacity (capacity of leucocyte plasma membrane; fluorescence of at least one bead currently attached to the membrane), and phagocytic efficiency (percentage of leucocytes that had engulfed microspheres; fluorescence of at least three beads).

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c. Statistical analysis

Statistical analysis was performed with the software R v.2.14.1 (R Core Team, 2014). Normality was checked using Shapiro's tests and homogeneity of variance was tested using Levene's test ($p \le 0.05$). When the normality assumption was not fulfilled, data were logtransformed if this improved normality.

209 To evaluate differences in biomarker levels between "experimental levels", three different210 statistical analyses were made:

First, at each sampling period, the differences between *laboratory* and *mesocosm* or *field* data were tested either by an ANCOVA followed by a Dunnett test (parametric data, $p \le 0.05$) or by a Kruskal-Wallis test followed by a Dunn test with comparison to the control (non-parametric data, $p \le 0.05$, "dunn.test.control" function of R package "PMCMR").

To determine whether *mesocosm* or *field* data fitted within the *laboratory* reference
ranges, the percentage of data outside the reference ranges were calculated for each

218 condition. Unilateral binomial tests ($p \le 0.05$) were performed at each sampling 219 period to detect if there was a significant increase of data out of the ranges between 220 *laboratory* data and *mesocosm* or *field* data.

The predictive power of the previously-built models, which were based on data
 collected in *laboratory* conditions, towards *mesocosm* and *field* data was also
 evaluated using the predictive squared correlation coefficient (Q²) as previously
 described (Marchand *et al.* 2019).

The differences in biomarker levels between the three "experimental levels" highlighted the need for specific *field* reference ranges. These were determined with the following steps:

First, a linear model that included all confounding factors (period, experimental
level, sex and size) and their binary and ternary interactions was adjusted for each biomarker.
Period, experimental level and sex were included as qualitative variables whereas size was
included as a quantitative variable. Though the aim was to build field reference ranges, data
obtained in mesocosms were included in order to consolidate confounding effects if
appropriate, for example by confirming relationships between size and biomarker levels.

- The effects of each confounding factor were evaluated using a type II analysis of covariance (ANCOVA) (Shaw and Mitchell-Olds, 1993) ("Anova" function of R package "car", $p \le 0.05$), owing for the fact the data was slightly unbalanced due to postmortem identification of sex. The ANCOVA model for each biomarker was selected using a downwards stepwise procedure based on the AIC and then by removing effects that were not statistically significant according to the F-test, starting from the highest order interactions. Reference ranges were estimated using the 95% prediction interval. The model was evaluated using the coefficient of determination (R²), and the
 predictive squared correlation coefficient determined by Leave-One-Out (Q²LOO).

242 III. Results

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a. Normality and heteroscedasticity

The results of normality tests (Shapiro test) and heteroscedasticity tests (Levene's test) made 244 on residuals of the linear models were presented in the Table II. Briefly, after log-245 transformation of biomarker values, the normality conditions were complete in most of the 246 groups. Nevertheless, some deviations from the normality assumption were found at specific 247 248 periods, mainly for the respiratory burst index (Table II). Levene's test revealed differences in variance between groups mainly for respiratory burst index, capacity and phagocytosis 249 efficiency (Table II). Observation of the variance in each group showed higher variances in 250 251 laboratory data regarding respiratory burst index in June, capacity in February, and phagocytosis efficiency in December (data not shown). 252

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b. Comparison between mesocosm and laboratory data

The results of the comparisons of means between "experimental levels" (Table III) were compared to results based on percentages of *mesocosm* data outside the *laboratory* reference ranges (Figure 1): the two methods provided complementary results for all biomarkers.

The results presented hereafter are based on statistical significance levels of 5%. Since the calculating Q² were low for granulocyte-macrophage (-1,54) and apoptotic cell (-1,07) percentages, the models predict some discrepancy between *laboratory* and *mesocosm* data in December (40% for apoptotic cell percentage, binomial test: $p = 2.86 \times 10^{-6}$; 20 % for granulocyte-macrophage percentage, binomial test: p = 0.0159; Figure 1) and in June (27.8 % for apoptotic cell percentage, binomial test: p = 0.0155; Figure 1, Table IV) whereas the comparison of means test does not (Table III). Only the granulocyte-macrophage percentage difference observed in June, with drastically highest values in *mesocosm* condition was observed with both methods (88.8 % of values out of range, binomial test: $p = 1.31 \times 10^{-21}$, Figure 1; Dunn test, $p = 3.2 \times 10^{-8}$, Table III).

A better predictivity of the models was shown for phagocytic activity, including efficiency 267 $(Q^2 = -0.40)$ and capacity $(Q^2 = -0.36)$, and necrotic cell percentage $(Q^2 = -0.22)$. The 268 significant high decrease of phagocytic efficiency shown in February (Dunnett test: p = 1.71 x 269 10^{-6} , Table III) and June (Dunnett test: $p = 1.71 \times 10^{-6}$, Table III) were also detected with the 270 predictive model, with 25 % (binomial test: $p = 2.57 \times 10^{-3}$) and 27.8 % (binomial test: p =271 272 0.00155) of deviation respectively (Figure 1). In the same way, no significant impact was shown with both methods in August and October. On the other hand, in December, the 273 predictive model shown dissimilarity between *laboratory* and *mesocosm* efficiency 274 percentages (60% of deviation, binomial test: $p = 2.11 \times 10^{-11}$; Figure 1) without statistical 275 significance (Table III). On the opposite, in April, statistical test shown a significant light 276 decrease of efficiency percentage in *mesocosm* condition (Dunnett test: $p = 8.5 \times 10^{-18}$, Table 277 III) without increase of error percentage with the model (Figure 1). Concerning the 278 phagocytic capacity, a good projection of mesocosm data on laboratory reference ranges was 279 observed in February, April, August and October (Figure 1). Nevertheless, the significant 280 light modulation of capacity percentage observed in December (Dunnett test: $p = 1 \times 10^{-6}$, 281 light increase in *mesocosm* condition, Table III) and June (Dunnett test: $p = 1 \times 10^{-6}$, light 282 decrease in *mesocosm* condition, Table III) was not detected by the predictive model (10%, 283 binomial test: p = 0.264, and 0%, respectively). The necrotic cell percentage present a quite 284 good projection of mesocosm data on laboratory reference range compared to statistical test, 285 286 with significant highest values in *mesocosm* condition detected in June using both methods (77.7 % of deviation, binomial test: $p = 2.12 \times 10^{-19}$; Dunn test, $p = 3.2 \times 10^{-8}$). Nevertheless, 287

in October, the significantly low increase detected with Dunn test (p = 0.0021, Table III) was not observed using projection on *laboratory* reference ranges (0 % of deviation, Figure 1).

Moreover, positive calculating Q^2 values were shown concerning respiratory burst index (Q^2 = 290 0.03) and lysosomal presence ($Q^2 = 0.17$) suggesting better projection deviations for these 291 biomarker models. Lower values of respiratory burst index were found in mesocosm 292 compared to laboratory experimental level at June period with both methods (27.7 % of 293 deviation, binomial test, p = 0.0001; Dunn test, $p = 1.3 \times 10^{-6}$, Table III). Even if the 294 lysosomal presence in *mesocosm* was slightly lower in December (Dunnett test p = 0.001) and 295 slightly higher in February (Dunn test, p = 0.00019) and in April (Dunn test, p = 0.034) in 296 297 mesocosm compared to laboratory data (Table III), the projection of mesocosm data on laboratory reference ranges was good for this biomarker (Figure 1). 298

The more important deviations between *mesocosm* and *laboratory* data were observed in June. One the other hand, in August no significant differences were identified whatever the immunomarker tested and the method used.

302 Overall, for all immunomarkers except lysosomal presence, there were statistically significant differences in seasonal trends between mesocosm and laboratory data (F-test between nested 303 regression models, p<0.05). When data collected in June was removed from the analysis, no 304 statistically significant differences were identified either regarding necrotic cell percentage. 305 The proportion of total variance explained by period and all related interactions is compared 306 307 to the proportion of total variance explained by the differences in seasonal trend in Table VI. Seasonal trends were strongest for phagocytosis efficiency and although differences in 308 seasonal trend between mesocosm and laboratory data were comparatively weak (5.54 % of 309 total variance), they represented significant contributions to the model (F-test, $p=6.44 \times 10^{-11}$). 310 Differences in seasonal trends were also particularly marked for necrotic cell percentage (due 311 to June, 19.2 % of total inertia, F-test, $p=5.86 \times 10^{-14}$) and macrophage-granulocyte 312

percentage (18.4 % of total inertia, p =6.66 x 10^{-16}). Lysosomal presence, for which no differences in seasonal trend were noted between mesocosm and laboratory data, was the immunomarker with the weakest seasonal trend in mesocosm and laboratory data (26.5 % of total inertia).

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c. Comparison between *field* and *laboratory* data

Globally, projection of *field* data on *laboratory* reference ranges revealed high discrepancy to
the model predictions (Figure 1). Like for *mesocosm* data, when these results were compared
to classic statistical analysis (Table III), some dissimilarities were found.

321 With a Q^2 value highly negative, equal to -5.26, the phagocytosis efficiency was the biomarker that fit the less to the *laboratory* model. Indeed, differences between *laboratory* 322 and *field* experimental conditions were found at each period, except in December, with higher 323 values for *field* data in August (Dunnett test: : $p = 1 \times 10^{-7}$), but lower values in February 324 (Dunn test: $p = 7.11 \times 10^{-15}$), April (Dunnett test: $p = 1.51 \times 10^{-18}$) and October (Dunn test: p325 = 1.51×10^{-18}) (Table III). These differences were observable by projection of *field* data on 326 *laboratory* reference ranges in February (85 % of deviation, binomial test: $p = 7.52 \times 10^{-20}$) 327 and April (65 % of deviation, binomial test: $p = 6.79 \times 10^{-13}$) but not in August (5 % of 328 329 deviation, binomial test: : p = 1 and October (10 % of deviation, binomial test: : p = 2.64). Moreover, the model predicts a difference in the December period (27.8 % of deviation, 330 binomial test: p = 0.00155), not detected by statistical tests (Figure 1). 331

Low predictive power was also found for the granulocyte-macrophage percentage model (Q² = - 1.1), for the respiratory burst index model (Q² = - 0.72), for the phagocytic capacity model (Q² = - 0.51), for the necrotic cell percentage (Q² = - 0.44), and for apoptotic cell percentage (Q² = 0.016). The significant increase of granulocyte-macrophage percentages shown in December (Dunnett test: p = 0.01, Table III), February (Dunn test: $p = 1 \ge 10^{-4}$, Table III) and

October (Dunn test: $p = 1 \ge 10^{-6}$, Table III) were also detected with the predictive model, with 337 33.3 % (binomial test: $p = 1.72 \times 10^{-4}$), 50 % (binomial test: $p = 1.13 \times 10^{-8}$) and 25 % 338 (binomial test: : $p = 2.57 \times 10^{-3}$) of deviation, respectively (Figure 1). On the contrary, the 339 significant increase in April (Dunn test: $p = 1 \ge 10^{-6}$) was not detected by the statistical model 340 (5 % of deviation, binomial test: p = 1.00). Even if the respiratory burst index was higher in 341 February (Dunn test: $p = 2.1 \times 10^{-4}$), April (Dunn test: p = 0.00015) and October (Dunn test: p342 = 3.8 x 10⁻⁵), but lower in August (Dunn test: $p = 2.95 \times 10^{-5}$) in *field* compared to *laboratory* 343 data (Table III), the proportion of values out of range was not significantly different from the 344 expected 5%. As it was detected by the statistical model (100 % of values out of range, 345 binomial test: $p = 9.54 \times 10^{-27}$), strong decrease of capacity was observed in August (Dunnett 346 test: $p = 1 \times 10^{-6}$). On the contrary, the low capacity decreases observed in February (Dunn 347 test: $p = 4 \times 10^{-5}$) and in October (Dunnett test: $p = 5 \times 10^{-4}$) was not detected by data 348 projection on reference ranges (15 % of deviation and 0 % of deviation respectively; Figure 349 1). For the necrotic cell percentage, the same effects were detected using both methods in 350 February (Dunnett test: $p = 1.5 \times 10^{-6}$; 40 % of deviation, binomial test: $p = 2.86 \times 10^{-6}$) and 351 August (Dunn test: $p = 3 \times 10^{-4}$) but not in December (Table III). Considering apoptotic cell 352 percentage, significant higher values were found in *field* experimental condition in April 353 (Dunnett test: $p = 2 \times 10^{-4}$) that matched with the model predictions (35 % of deviation; 354 355 binomial test: $p = 3.39 \times 10^{-5}$).

With a Q² equal to 0.23, the model designed for the lysosomal presence was the most predictive for *field* data. Compared to *laboratory* data, lysosomal presence was significantly higher in *field* experimental condition in February (Dunnett test: $p = 1 \ge 10^{-4}$) and April (Dunnett test: $p = 2.7 \ge 10^{-16}$) but lower in August (Dunnett test: $p = 1 \ge 10^{-5}$) and October (Dunnett test: p = 0.02; Table III). These effects were found again using projection on reference ranges in February (25 % of values out of range; binomial test: p = 0.002), April 80 362 % of values out of range; binomial test: $p < 2.2 \ge 10^{-11}$ and August (50 % of values out of 363 range; binomial test: $p = 1.13 \ge 10^{-8}$) but not in October (Figure 1).

Statistically significant differences in seasonal trends between field and laboratory data for all immunomarkers (F-test between nested regression models, p<0.05). Seasonal trends were strongest for lysosomal presence, capacity, and phagocytosis efficiency (F-test, p<10⁻¹⁶), for which differences in seasonal trends between laboratory and field data represented 26.6 %, 23.8 %, and 18.8 % of total variance respectively. Differences in seasonal trends were less strong for macrophage-granulocyte percentage (3.09 % of total inertia, p =2.17 x 10⁻²).

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d. Evaluation of the impact of the "experimental level" factor on the proposed reference ranges

373 The models designed for respiratory burst index and for apoptotic cells explained only a small 374 proportion of the data variability with R² equal to 0.29 and 0.33 respectively. The necrotic 375 cells percentage was slightly better ($R^2 = 0.43$) explained by the confounding factors ("month", "experimental level", "sex", "size") and their interactions (Table IV). The global 376 type-II ANCOVA revealed that "month", "experimental level" (for apoptotic cell percentages, 377 378 respiratory burst and necrotic cell percentages) and "size" (for respiratory burst index, apoptotic cell percentages, and necrotic cell percentages) all had a statistically significant 379 effect on these immunomarkers. In addition to the "month:experimental level" interaction 380 which influenced these three immunomarkers, the apoptotic cell percentages were also 381 impacted by "month:size" interaction and the necrotic cell percentages by "month:sex" 382 383 interaction (Table IV). The predictivity of the models including all the significant factors and interactions was relatively low for the immunomarkers cited above. With a O²LOO almost 384 equal, the apoptotic cell percentages ($Q^2LOO = 0.23$) and the respiratory burst index (Q^2LOO 385

= 0.24) were the less well predicted biomarkers by the model. Higher, but still low, predictive power was found for the necrotic cell percentages, with a Q²LOO equal to 0.43 (Table VI).

Granulocyte-macrophage percentages ($R^2 = 0.52$), lysosomal presence ($R^2 = 0.54$), and 388 389 capacity ($R^2 = 0.63$) presented better correlation with these factors (Table IV). Indeed, the "month", the 'experimental level" and their interaction have a great influence on these three 390 391 immunomarkers. With granulocyte-macrophage percentages and capacity the interaction "period:size" was strongly significant. Effects of "size" were more considerable for capacity 392 than for granulocyte-macrophage percentages and lysosomal presence. Granulocyte-393 macrophage percentages were also influenced by "period:sex" and "experimental level:size" 394 395 interactions (Table IV). Medium-high level of predictivity was found for theses markers, with Q²LOO equal to 0.51 for lysosomal presence, 0.60 for phagocytic capacity and 0.46 for 396 granulocyte-macrophage percentages. 397

The phagocytic efficiency was highly correlated to the factors studied here ($R^2 = 0.80$). The "period" and the "experimental level" factors have a strong influence while the "size" factor influence was less important (Table IV). This marker was also highly influenced by "period:size", "period:experimental level" and "size:experimental level" interactions. This good correlation with confounding factors was associated with a high predictive power of the model including significant factors and interactions (Q²LOO = 0.77) (Table IV).

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405 IV. Discussion

Use of robust reference values for biomarkers is important in biomonitoring programs to minimize data misinterpretation. Nevertheless, due to the influence of environmental factors on biomarkers (Hanson et al., 2010; Jolly et al., 2012; Sanchez and Porcher, 2009), the addition of confounding factors in the statistical model should improve predictions of

biomarker reference values in function of site characteristics (Coulaud et al., 2011; Maltby et 410 al., 2002). In this way, many authors developed laboratory reference values comprising 411 confounding factor effect for field evaluation (Barrick et al., 2018, 2016; Burgeot et al., 2010; 412 413 Coulaud et al., 2011; Krell et al., 2011; Maltby et al., 2002). In case of stickleback innate immunomarkers, the first step was performed in a previous study (Marchand et al., 2019) by 414 assessing the effect of sampling period, fish size and sex in laboratory conditions. 415 Nonetheless, before using these reference values on the field, the variations due to natural 416 417 conditions should be included. Natural modulation of the immune response could be due to variations of environmental conditions, especially temperature and photoperiod (Bly and 418 Clem, 1992; Bowden et al., 2007; Dittmar et al., 2014; Morgan et al., 2008; Porter et al., 419 2001) and to reproductive status (Hou et al., 1999; Kortet et al., 2003; Slater and Schreck, 420 1998; Zapata et al., 1992). During the one-year sampling, temperature and photoperiod 421 422 variations were similar in all three experimental conditions and the influence of sexual hormones during breeding period was expected to be similar. 423

424 To determine if *laboratory* reference values could predict other experimental levels, new data was compared to data obtained in laboratory conditions either using reference ranges obtained 425 with a regression model or by simply comparing means at given periods. The results using 426 both methodologies differed for several biomarkers. The laboratory reference ranges 427 developed previously took the effect of fish size into account on most of these 428 immunomarkers, except respiratory burst index and lysosomal presence (Marchand et al. 429 430 2019) for which the models developed in this paper are therefore not expected to be an improvement over comparisons of means at a given period. For the other five 431 immunomarkers, since fish body size did not cover the same range in mesocosm and field data 432 as in the laboratory data, statistically significant differences in means could actually be due to 433 434 the confounding effect of fish size. For example, the efficiency possessed a strong "size"

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effect in December and only the comparison with the *laboratory* reference ranges showed
differences between experimental levels. However, this strong effect of size in December was
based on laboratory data where two distinct fish sizes were analysed. Thus, this point had to
be confirmed with additional data. Sex had also been identified as a confounding factor of
necrotic cell percentage and granulocyte-macrophage distribution values.

Integration of more data in the regression models of immunomarkers can strengthen the 440 441 knowledge on the effect of confounding effects of size and sex, even when the new data is not obtained in the same conditions. First, the immunomarkers that had been log-transformed in 442 443 the first study were again log-transformed in the present study. Secondly, the additional data 444 obtained in mesocosm and field conditions also confirmed that including sex was only relevant for necrotic cell percentage and granulocyte-macrophage distribution. No significant 445 interactions between sex and experimental conditions were observed, which tends to confirm 446 that the effect of sex is independent of the experimental conditions. Thirdly, size had not been 447 included in the regression models based on laboratory data for respiratory burst index and 448 449 lysosomal presence, but when mesocosm and field data were added, a linear effect of size on respiratory burst index (see Figure 2 in supplementary data) improved model quality, 450 independently of experimental conditions and sampling period. The additional data did not 451 452 however confirm the strong effect of size on efficiency in December, which appeared to be specific to the set of data obtained in laboratory. 453

Even so, regarding percentages of data outside *laboratory* reference values, Dunn/Dunnett statistical tests, and Q², the *laboratory* reference ranges developed in the precedent work (Marchand et al., 2019) does not fit perfectly with *mesocosm* and even less with *field* data. Seasonal trends varied significantly between experimental levels. Furthermore, although lysosomal presence showed little seasonal variation between mesocosm and laboratory data, large differences were observed in field conditions. Conversely, although the seasonal trends

in macrophage-granulocyte percentages were similar in field and laboratory conditions, they 460 were different in mesocosms. All these statistical aspects demonstrated that the "experimental 461 level" factor have a strong effect on all the studied immunomarkers. Furthermore, as 462 463 demonstrated by the "experimental level:period" interaction, the "experimental level" factor was also dependent of the season for almost all immunomarkers which implies that seasonal 464 variations were not the same in the various experimental conditions. As previously discussed, 465 the immune response could be influenced by many abiotic (Bowden, 2008) and biotic (Kortet 466 et al., 2003; Robertson et al., 2016) factors, which explain this discrepancy between 467 experimental conditions. Since the fish used in the different experimental condition originated 468 from the same regional ecotype and the temperature and photoperiod were quite similar, 469 differences in pathogen diversity and abundance were likely causes. In fact, pathogen 470 presence induces specific adaptations which may lead to differences in immune profiles of 471 472 populations living in separate environments (Scharsack et al., 2016). Nevertheless, even if this specific adaptation of fish to their pathogenic environment can be responsible for a decrease 473 474 in immunity variability (Hablützel et al., 2016), the ANCOVA regression model relies on the assumption that residual variability is the same in each condition. The differences in 475 variability observed at specific periods for some immunomarkers imply that at some periods, 476 the reference ranges will be narrower (e.g. respiratory burst index in June) than expected 477 given the variability at that period, and in some cases, they will be slightly larger (e.g. 478 respiratory burst index at periods other than June). Food allocation and intestinal microbiota 479 also impact the fish immune response (Gómez and Balcázar, 2008; Kosiewicz et al., 2014; 480 Waagbo, 1994). In laboratory and mesocosm conditions, fish were daily fed with bloodworms 481 unlike in *field* conditions where the food may have been more diversified. This parameter may 482 also be an explicative factor in the differences between conditions. Another factor fixed in 483 laboratory experimental level but not in mesocosm and field experimental levels was the fish 484

485 age. Since this factor was able to modulate fish immunity, in both innate and acquired486 component, they should be considered (Robertson et al., 2016).

As for laboratory reference ranges (Marchand et al. 2019), the newly created models 487 including the three type of data (laboratory, mesocosm and field) integrated especially the 488 "size" effect. In fact, each immunomarker were differently influenced by the "size", with 489 490 more impact on necrotic cell percentages and capacity and no statistically significant effect 491 shown on respiratory burst index and lysosomal presence. Likewise, as shown by the coefficient of determination (R²) of the models, the part of data variability explained by these 492 493 confounding factors was also variable between immunomarkers. Thus, whereas the efficiency, 494 the capacity and the granulocyte-macrophage percentages were well-explained by the considered confounding factors, the respiratory burst index and the apoptotic cell percentages 495 were poorly explained. In some cases, high values of the coefficient of determination are due 496 to large effects of experimental conditions. This can be observed for example with lysosomal 497 presence: in our previous paper the R² reported was 13 % (Marchand et al., 2019), whereas it 498 499 is 53 % in the model including all data. The total variability due to experimental level when 500 data from mesocosms or data from field conditions were compared to laboratory conditions is 501 reported in Table VI: when data from field and laboratory are compared, as much as 40.8~%502 of variability is due to the experimental level (capacity). The importance of experimental level on immunomarkers may hinder extrapolation to other field conditions, since our experiments 503 were designed to limit differences in photoperiod, temperature and fish population 504 505 characteristics. Further research is need in similar filed conditions with the least possible contaminations. The fact that, for some immunomarkers, confounding factors only have a 506 small effect on an immunomarker, as for respiratory burst index with only 28 % of data 507 508 variability explained by the four confounding factors, can have several causes and 509 implications. This may indicate either high inter-individual variability, high measurement 510 uncertainty, unidentified inter-individual confounding factors such as reproductive or health status, or a lack of sensitivity to environmental factors. At the present, no studies have 511 established a relationship between sensitivity to environmental confounding factors and 512 sensitivity to various types and levels of pollution: further research is needed to compare 513 values of immunomarkers in similar environmental conditions with different levels of 514 pollution in a biomonitoring context. Depending on the sensitivity of these immunomarkers to 515 various types of contamination, the *field* reference ranges derived in these models could help 516 517 to define robust reference values useful for large scale biomonitoring programs (Maltby et al., 2002). 518

519 V. Conclusions

This study focused on the evaluation of laboratory-based reference ranges, created on three-520 521 spined stickleback immunomarkers, for mesocosm and field conditions. The "experimental level" factor influences in a seasonal-dependent pattern all the studied immunomarkers, 522 523 therefore care must be taken when extrapolating reference values or ranges between different experimental conditions. Finally, this study argued for considering the influence of specific 524 field confounding factors. In this way, field reference ranges were developed. Nevertheless, 525 526 before using them for large scale biomonitoring programs, these *field* reference ranges should be tested on multiple uncontaminated and contaminated sites in order to identify the 527 biomarkers which offer the best compromise between sensitivity to contamination, and either 528 reproducibility in similar conditions or well-characterized effects of abiotic confounding 529 factors. 530

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Bibliography

- Ahmad, I., M. Pacheco, and M. A. Santos. 2004. Enzymatic and nonenzymatic antioxidants as an adaptation to phagocyte-induced damage in Anguilla anguilla L. following in situ harbor water exposure. Ecotoxicology and Environmental Safety 57:290-302.
- Ambrozova, G., Pekarova, M., Lojek, A., 2011. The effect of lipid peroxidation products on reactive oxygen species formation and nitric oxide production in lipopolysaccharidestimulated RAW 264.7 macrophages. Toxicol. Vitr. 25, 145–152. https://doi.org/10.1016/j.tiv.2010.10.006.
- Ansar Ahmed, S. 2000. The immune system as a potential target for environmental estrogens (endocrine disrupters): a new emerging field. Toxicology 150:191-206.
- Arkoosh, M. R., E. Casillas, E. Clemons, A. N. Kagley, R. Olson, P. Reno, and J. E. Stein.1998. Effect of Pollution on Fish Diseases: Potential Impacts on Salmonid Populations.Journal of Aquatic Animal Health 10:182-190.
- Arnold, H., H.-J. Pluta, and T. Braunbeck. 1995. Simultaneous exposure of fish to endosulfan and disulfoton in vivo: ultrastructural, stereological and biochemical reactions in hepatocytes of male rainbow trout (Oncorhynchus mykiss). Aquatic Toxicology 33:17-43.
- Bado-Nilles, A., Betoulle, S., Geffard, A., Gagnaire, B., Porcher, J.M., Sanchez, W., 2013. Flow cytometry detection of lysosomal membrane integrity in the three spined stickleback (Gasterosteus aculeatus L.) immune cells: applications in environmental aquatic immunotoxicology. Comp. Biochem. Physiol. Part A Mol. Integr. Physiol. 163, S39. https://doi.org/10.1016/j.cbpa.2012.05.115

Bado-Nilles, A., Jolly, S., Porcher, J.-M., Palluel, O., Geffarda, A., Gagnaire, B., Betoulle, S.,

Wilfried, S., 2014. Applications in environmental risk assessment of leucocyte apoptosis, necrosis and respiratory burst analysis on the European bullhead, Cottus sp. Environ. Pollut. 184, 9–17.

- Bado-Nilles, A., C. Quentel, M. Auffret, S. Le Floch, T. Renault, and H. Thomas-Guyon.
 2009. Immune effects of HFO on European sea bass, Dicentrarchus labrax, and Pacific oyster, Crassostrea gigas. Ecotoxicology and Environmental Safety 72:1446-1454
- Barrick, A., Châtel, A., Marion, J.M., Perrein-Ettajani, H., Bruneau, M., Mouneyrac, C., 2016. A novel methodology for the determination of biomarker baseline levels in the marine polychaete Hediste diversicolor. Mar. Pollut. Bull. 108, 275–280. https://doi.org/10.1016/j.marpolbul.2016.04.056
- Barrick, A., Marion, J.-M., Perrein-Ettajani, H., Châtel, A., Mouneyrac, C., 2018. Baseline levels of biochemical biomarkers in the endobenthic ragworm Hediste diversicolor as useful tools in biological monitoring of estuaries under anthropogenic pressure. Mar. Pollut. Bull. 129, 81–85. https://doi.org/10.1016/j.marpolbul.2018.02.006Benchalgo, N., F. Gagné, and M. Fournier. 2014. Immunotoxic effects of an industrial waste incineration site on groundwater in rainbow trout (Oncorhynchus mykiss). Journal of Environmental Sciences 26:981-990.
- Betoulle, S., C. Duchiron, and P. Deschaux. 2000. Lindane increases in vitro respiratory burst activity and intracellular calcium levels in rainbow trout (Oncorhynchus mykiss) head kidney phagocytes. Aquatic Toxicology 48:211-221.
- Bly, J.E., Clem, L.W., 1992. Temperature and teleost immune functions. Fish Shellfish Immunol. 2, 159–171. https://doi.org/10.1016/S1050-4648(05)80056-7
- Bols, N.C., Brubacher, J.L., Ganassin, R.C., Lee, L.E.J., 2001. Ecotoxicology and innate immunity in fish. Dev. Comp. Immunol. 25, 853–873. https://doi.org/10.1016/S0145-

- Bowden, T.J., 2008. Modulation of the immune system of fish by their environment. Fish Shellfish Immunol. 25, 373–383. https://doi.org/10.1016/j.fsi.2008.03.017
- Bowden, T.J., Thompson, K.D., Morgan, A.L., Gratacap, R.M.L., Nikoskelainen, S., 2007.
 Seasonal variation and the immune response: A fish perspective. Fish Shellfish Immunol.
 22, 695–706. https://doi.org/10.1016/j.fsi.2006.08.016
- Burgeot, T., Gagné, F., Forget-leray, J., Bocquené, G., 2010. Acethylcholinesterase: Methodology development of a biomarker and challenges of its application for biomonitoring. Aquat. Living Resour. 17, 309–316.Cabas, I., E. Chaves-Pozo, V. Mulero, and A. Garcia-Ayala. 2018. Role of estrogens in fish immunity with special emphasis on GPER1. Dev Comp Immunol 89:102-110.
- Chadzinska, M., Tertil, E., Kepka, M., Hermsen, T., Scheer, M., Lidy Verburg-van Kemenade, B.M., 2012. Adrenergic regulation of the innate immune response in common carp (Cyprinus carpio L.). Dev. Comp. Immunol. 36, 306–316. https://doi.org/10.1016/j.dci.2011.04.010
- Chilmonczyk, S., Monge, D., 1999. Flow cytometry as a tool for assessment of the fish cellular immune response to pathogens. Fish Shellfish Immunol. 9, 319–333. https://doi.org/10.1006/fsim.1998.0188
- Coulaud, R., Quéau, H., Garric, J., Charles, S., Bernard, C., Lyon, U., 2011. In situ feeding assay with Gammarus fossarum (Crustacea): Modelling the influence of confounding factors to improve water quality biomonitoring. Water Res. 45, 6417–6429.
- de Kermoysan, G., S. Joachim, P. Baudoin, M. Lonjaret, C. Tebby, F. Lesaulnier, F. Lestremau, C. Chatellier, Z. Akrour, E. Pheron, J.-M. Porcher, A. R. R. Péry, and R.

Beaudouin. 2013. Effects of bisphenol A on different trophic levels in a lotic experimental ecosystem. Aquatic Toxicology 144-145:186-198.Danion, M., S. Le Floch, R. Kanan, F. Lamour, and C. Quentel. 2011. Effects of in vivo chronic hydrocarbons pollution on sanitary status and immune system in sea bass (Dicentrarchus labrax L.). Aquatic Toxicology 105:300-311.

Dittmar, J., Janssen, H., Kuske, A., Kurtz, J., Scharsack, J.P., 2014. Heat and immunity: An experimental heat wave alters immune functions in three-spined sticklebacks (Gasterosteus aculeatus). J. Anim. Ecol. 83, 744–757. https://doi.org/10.1111/1365-2656.12175

Ellis, A.E., 1999. Immunity to bacteria in fish. Fish Shellfish Immunol. 9, 291–308.

- Flammarion, P., Devaux, A., Garric, E.J., 2000. Marqueurs biochimiques de pollution dans les écosystèmes aquatiques continentaux. Exemple d'utilisation prospectives pour le gestionnaire. Bull. Fr. Pêche Piscic. Bull. Fr. Pêche Piscic 3571358358. https://doi.org/10.1051/kmae/2001045
- Gagnaire, B., Thomas-Guyon, H., Renault, T., 2004. In vitro effects of cadmium and mercury on Pacific oyster, Crassostrea gigas (Thunberg), haemocytes. Fish Shellfish Immunol. 16, 501–512. https://doi.org/10.1016/j.fsi.2003.08.007
- Gómez, G.D., Balcázar, J.L., 2008. A review on the interactions between gut microbiota and innate immunity of fish: Table 1. FEMS Immunol. Med. Microbiol. 52, 145–154. https://doi.org/10.1111/j.1574-695X.2007.00343.x
- Guicciardi, M. E., M. Leist, and G. J. Gores. 2004. Lysosomes in cell death. Oncogene 23:2881-2890.

Hablützel, P.I., Brown, M., Friberg, I.M., Jackson, J.A., 2016. Changing expression of

vertebrate immunity genes in an anthropogenic environment: a controlled experiment. BMC Evol. Biol. 16, 175. https://doi.org/10.1186/s12862-016-0751-8

- Hanson, N., 2011. Using biological data from field studies with multiple reference sites as a basis for environmental management: The risks for false positives and false negatives. J. Environ. Manage. 92, 610–619. https://doi.org/10.1016/j.jenvman.2010.09.026
- Hanson, N., Förlin, L., Larsson, Å., 2010. Spatial and annual variation to define the normal range of biological endpoints: An example with biomarkers in perch. Environ. Toxicol. Chem. 29, 2616–2624. https://doi.org/10.1002/etc.300
- Holtzman, E. 1989. Lysosomes. Plenum Press, New York.
- Hou, Y., Suzuki, Y., Aida, K., 1999. Changes in Immunoglobulin Producing Cells in Response to Gonadal Maturation in Rainbow Trout. Fisk. Sci. 65, 844–849.
- Jolly, S., Bado-Nilles, A., Lamand, F., Turies, C., Chadili, E., Porcher, J.M., Betoulle, S., Sanchez, W., 2012. Multi-biomarker approach in wild European bullhead, Cottus sp., exposed to agricultural and urban environmental pressures: Practical recommendations for experimental design. Chemosphere 87, 675–683.
- Jones, L. O., and J. L. Stafford. 2019. Imaging flow cytometry and confocal microscopybased examination of F-actin and phosphoinositide dynamics during leukocyte immunetype receptor-mediated phagocytic events. Developmental & Comparative Immunology 92:199-211.
- Kortet, R., Taskinen, J., Sinisalo, T., Jokinen, I., 2003. Breeding-related seasonal changes in immunocompetence, health state and condition of the cyprinid fish, Rutilus rutilus, L. Biol. J. Linn. Soc. 78, 117–127. https://doi.org/10.1046/j.1095-8312.2003.00136.x

Kosiewicz, M.M., Dryden, G.W., Chhabra, A., Alard, P., 2014. Relationship between gut

microbiota and development of T cell associated disease. FEBS Lett.

- Krell, B., Moreira-Santos, M., Ribeiro, R., 2011. An estuarine mudsnail in situ toxicity assay based on postexposure feeding. Environ. Toxicol. Chem. 30, 1935–1942. https://doi.org/10.1002/etc.585
- Kreutz, L. C., L. J. Gil Barcellos, A. Marteninghe, E. Davi dos Santos, and R. Zanatta. 2010. Exposure to sublethal concentration of glyphosate or atrazine-based herbicides alters the phagocytic function and increases the susceptibility of silver catfish fingerlings (Rhamdia quelen) to Aeromonas hydrophila challenge. Fish & Shellfish Immunology 29:694-697.
- Krumschnabel, G., C. Manzl, C. Berger, and B. Hofer. 2005. Oxidative stress, mitochondrial permeability transition, and cell death in Cu-exposed trout hepatocytes. Toxicology and Applied Pharmacology 209:62-73.
- Kumari, J., Sahoo, P.K., Swain, T., Sahoo, S.K., Sahu, a. K., Mohanty, B.R., 2006. Seasonal variation in the innate immune parameters of the Asian catfish Clarias batrachus. Aquaculture 252, 121–127. https://doi.org/10.1016/j.aquaculture.2005.07.025
- Kurz, T., A. Terman, B. Gustafsson, and U. T. Brunk. 2008. Lysosomes and oxidative stress in aging and apoptosis. Biochimica et Biophysica Acta (BBA) - General Subjects 1780:1291-1303.
- Lo, C. M., S. Morand, and R. Galzin. 1998. Parasite diversity/host age and size relationship in three coral-reef fishes from French Polynesia. Int J Parasitol 28:1695-1708.
- Magnadóttir, B., 2006. Innate immunity of fish (overview). Fish Shellfish Immunol. 20, 137– 151. https://doi.org/10.1016/j.fsi.2004.09.006

Maltby, L., Clayton, S.A., Wood, R.M., McLoughlin, N., 2002. Evaluation of the Gammarus

pulex in situ feeding assay as a biomonitor of water quality: Robustness, responsiveness, and relevance. Environ. Toxicol. Chem. 21, 361–368. https://doi.org/10.1002/etc.5620210219

- Marchand, A., J.-M. Porcher, C. Turies, E. Chadili, O. Palluel, P. Baudoin, S. Betoulle, and A. Bado-Nilles. 2017. Evaluation of chlorpyrifos effects, alone and combined with lipopolysaccharide stress, on DNA integrity and immune responses of the three-spined stickleback, Gasterosteus aculeatus. Ecotoxicology and Environmental Safety 145:333-339.
- Marchand, A., C. Tebby, R. Beaudouin, Y. M. I. Hani, J. M. Porcher, C. Turies, and A. Bado-Nilles. 2019. Modelling the effect of season, sex, and body size on the three-spined stickleback, Gasterosteus aculeatus, cellular innate immunomarkers: A proposition of laboratory reference ranges. Sci Total Environ 648:337-349.
- Misumi, I., A. T. Vella, J.-A. C. Leong, T. Nakanishi, and C. B. Schreck. 2005. p,p'-DDE depresses the immune competence of chinook salmon (Oncorhynchus tshawytscha) leukocytes. Fish & Shellfish Immunology 19:97-114.
- Mohanty, B.R., Sahoo, M., Sahoo, P.K., Mahapatra, K.D., Saha, J.N., 2014. Reference ranges and seasonal variations in innate immune responses of kalbasu , Labeo calbasu (Hamilton). Indian J. Fish 61, 57–62.
- Monserrat, J.M., Martínez, P.E., Geracitano, L. a., Lund Amado, L., Martinez Gaspar Martins, C., Lopes Leães Pinho, G., Soares Chaves, I., Ferreira-Cravo, M., Ventura-Lima, J., Bianchini, A., 2007. Pollution biomarkers in estuarine animals: Critical review and new perspectives. Comp. Biochem. Physiol. - C Toxicol. Pharmacol. 146, 221–234. https://doi.org/10.1016/j.cbpc.2006.08.012
- Morgan, A.L., Thompson, K.D., Auchinachie, N.A., Migaud, H., 2008. The effect of

seasonality on normal haematological and innate immune parameters of rainbow trout Oncorhynchus mykiss L. Fish Shellfish Immunol. 25, 791–799. https://doi.org/10.1016/j.fsi.2008.05.011

- Porter, M.J.R., Duncan, N., Handeland, S.O., 2001. Temperature, light intensity and plasma melatonin levels in juvenile Atlantic salmon. J. Fish Biol. 58, 431–438. https://doi.org/10.1006/jfbi.2000.1455
- Press, C., Evensen, Ø., 1999. The morphology of the immune system in teleost fishes. Fish Shellfish Immunol. 9, 309–318. https://doi.org/10.1006/fsim.1998.0181
- Reynaud, S., and P. Deschaux. 2006. The effects of polycyclic aromatic hydrocarbons on the immune system of fish: a review. Aquatic Toxicology 77:229-238.
- Risso-de Faverney, C., A. Devaux, M. Lafaurie, J. P. Girard, B. Bailly, and R. Rahmani. 2001. Cadmium induces apoptosis and genotoxicity in rainbow trout hepatocytes through generation of reactive oxygene species. Aquatic Toxicology 53:65-76.
- Robertson, S., Bradley, J.E., Maccoll, A.D.C., 2016. Measuring the immune system of the three-spined stickleback - investigating natural variation by quantifying immune expression in the laboratory and the wild. Mol. Ecol. Resour. 16, 701–713. https://doi.org/10.1111/1755-0998.12497
- Roméo, M., N. Bennani, M. Gnassia-Barelli, M. Lafaurie, and J. P. Girard. 2000. Cadmium and copper display different responses towards oxidative stress in the kidney of the sea bass Dicentrarchus labrax. Aquatic Toxicology 48:185-194.
- Salo, H. M., N. Hebert, C. Dautremepuits, P. Cejka, D. G. Cyr, and M. Fournier. 2007. Effects of Montreal municipal sewage effluents on immune responses of juvenile female rainbow trout (Oncorhynchus mykiss). Aquatic Toxicology 84:406-414.

- Santos, M. A., M. Pacheco, and I. Ahmad. 2006. Responses of European eel (Anguilla anguilla L.) circulating phagocytes to an in situ closed pulp mill effluent exposure and its association with organ-specific peroxidative damage. Chemosphere 63:794-801.
- Sanchez, W., Piccini, B., Ditche, J.M., Porcher, J.M., 2008. Assessment of seasonal variability of biomarkers in three-spined stickleback (Gasterosteus aculeatus L.) from a low contaminated stream: Implication for environmental biomonitoring. Environ. Int. 34, 791–798. https://doi.org/10.1016/j.envint.2008.01.005
- Sanchez, W., Porcher, J.-M., 2009. Fish biomarkers for environmental monitoring within the Water Framework Directive of the European Union. TrAC Trends Anal. Chem. 28, 150–158. https://doi.org/10.1016/j.trac.2008.10.012
- Scharsack, J.P., Franke, F., Erin, N.I., Kuske, A., B??scher, J., Stolz, H., Samonte, I.E., Kurtz, J., Kalbe, M., 2016. Effects of environmental variation on host???parasite interaction in three-spined sticklebacks (Gasterosteus aculeatus). Zoology 119, 375–383. https://doi.org/10.1016/j.zool.2016.05.008
- Shaw, R.G., Mitchell-Olds, T., 1993. Anova for Unbalanced Data: An Overview. Ecology 74, 1638–1645. https://doi.org/10.2307/1939922
- Shepherd, B. S., C. B. Rees, F. P. Binkowski, and F. W. Goetz. 2012. Characterization and evaluation of sex-specific expression of suppressors of cytokine signaling (SOCS)-1 and -3 in juvenile yellow perch (Perca flavescens) treated with lipopolysaccharide. Fish Shellfish Immunol 33:468-481.
- Slater, C. H., and C. B. Schreck. 1993. Testosterone alters the immune response of chinook salmon, Oncorhynchus tshawytscha. Gen Comp Endocrinol 89:291-298.
- Slater, C.H., Schreck, C.B., 1998. Season and physiological parameters modulate salmonid

leucocyte androgen receptor affinity and abundance. Fish Shellfish Immunol. 8, 379–391. https://doi.org/10.1006/fsim.1998.0150

- Tellez-Bañuelos, M. C., A. Santerre, J. Casas-Solis, A. Bravo-Cuellar, and G. Zaitseva. 2009. Oxidative stress in macrophages from spleen of Nile tilapia (Oreochromis niloticus) exposed to sublethal concentration of endosulfan. Fish & Shellfish Immunology 27:105-111.
- Uribe, C., Folch, H., Enriquez, R., Moran, G., 2011. Innate and adaptive immunity in teleost fish: a review. Vet. Med. (Praha). 56, 486–503.
- van der Oost, R., Beyer, J., Vermeulen, N.P.E., 2003. Fish bioaccumulation and biomarkers in environmental risk assessment : a review. Environ. Toxicol. Pharmacol. 13.
- Waagbo, R., 1994. The impact of nutritional factors on the immune system in Atlantic salmon, Salmo salar L.: a review. Aquac. Res. 25, 175–197. https://doi.org/10.1111/j.1365-2109.1994.tb00573.x
- Yamaguchi, T., H. Watanuki, and M. Sakai. 2001. Effects of estradiol, progesterone and testosterone on the function of carp, Cyprinus carpio, phagocytes in vitro. Comparative Biochemistry and Physiology Part C: Toxicology & Pharmacology 129:49-55.
- Zapata, a. G., Varas, a., Torroba, M., 1992. Seasonal variations in the immune system of lower vertebrates. Immunol. Today 13, 142–147. https://doi.org/10.1016/0167-5699(92)90112-K
- Zdolsek, J. M., G. M. Olsson, and U. T. Brunk. 1990. Photooxidative damage to lysosomes of cultures macrophages by acridine orange. Photochemistry and Photobiology 51:67-76.





Figure 1: Projection of *mesocosm* (black) and *field* (grey) data on the *laboratory* reference ranges. For apoptotic cells, respiratory burst index, lysosomal presence, phagocytosis capacity and efficiency, the lines represent the values predicted by the model and the prediction interval at 95 %. For necrotic cells percentage and granulocyte-macrophage percentage, the full lines represent prediction intervals for females and the dotted lines represent prediction intervals for males. With MFI =Mean Fluorescent Intensity and e.u = experimental units.

Table I: Description of the two data sets used to perform the comparison to *laboratory* reference ranges in this study. *Laboratory* data set used to create *laboratory* reference ranges was described in Marchand et al. (2019).

M	ESOCOSM		FIELD										
		Fish body size		n (M ; F)	Fish body size	Water parameters							
Sampling period	<i>n</i> (M ; F)	(mm)	Sampling period		(mm)	pН	Т (°С)	O2 (%)	Conductivity (µS/cm ²)				
October 2015	20 (7 ; 13)	45-57	October 2015	20 (8;12)	30 - 46	7.07	12	7.3	544				
December 2015	20 (12;8)	44-60	August 2017	20 (6; 14)	41 - 60	*	*	*	*				
February 2016	20 (9;11)	37-47	December 2017	20 (6 ; 14)	40 - 53	7.14	11.6	68.3	719				
April 2016	15 (6;9)	33-46	February 2017	20 (11;9)	41 - 60	7.42	11.3	75.2	706				
June 2016	18 (12;6)	40-51	April 2017	20 (10; 10)	42 - 52	7.7	13.9	79	799				
August 2016	19 (12 ; 7)	39-55											
TOTAL	112 (58 ; 54)	33-60	TOTAL	100 (41 ; 59)	30 - 60								

* Water parameters were not monitored at this period due to technical issues.

Table II: *p*-values from Shapiro tests inside each group and Levene tests for each immunomarker. Analysis were performed on residuals of the full linear model and biomarkers with * were log transformed before performing statistical tests.

							SH	APIR	0					
	LEVENE	Sov	Decemb	er	Februa	ry	April	l	June		August		October	
		Sex	Mesocosme	Field	Mesocosme	Field	Mesocosme	Field	Mesocosme	Field	Mesocosme	Field	Mesocosme	Field
A	4.54×10^{-2}	Female	0.53	0.39	0.74	0.72	0.33	0.14	0.87	-	0.80	0.26	0.63	0.07
Apoptotic cens*	4.34 X 10	Male	0.39	0.54	0.27	0.72	0.89	0.58	0.04	-	0.38	0.38	Octob Mesocosme 0.63 0.77 0.82 0.94 0.25 0.02 0.64 0.41 0.85 0.81 0.82 0.94	0.21
	2.54 10-2	Female	0.55	0.81	0.96	0.86	0.77	0.53	0.19	-	0.59	0.59	Octobe ield Mesocosme .26 0.63 .38 0.77 .59 0.82 .35 0.94 .22 0.25 .48 0.02 .17 0.64 .16 0.41 .73 0.85 .48 0.81 .59 0.82 .35 0.94 .48 0.81 .59 0.82 .35 0.94 .94 0.82 .49 0.06	0.30
Necrotic cells*	2.74x 10 ⁻²	Male	0.21	0.93	0.89	0.07	0.01	0.14	0.53	-	0.03	August Field August August </td <td>0.94</td> <td>0.37</td>	0.94	0.37
		Female	0.44	0.72	0.30	0.82	0.71	0.06	0.92	-	0.78	0.22	0.25	0.99
Granulocyte-macrophage	0.16	Male	0.20	0.41	0.91	0.41	0.74	0.03	0.86	-	0.64	0.48	0.02	0.28
Respiratory burst index*	LEVENE Sex December February Mesocosme Field Mesocosme <t< td=""><td>0.33</td><td>0.028</td><td>0.26</td><td>-</td><td>0.28</td><td>0.17</td><td>0.64</td><td>0.31</td></t<>	0.33	0.028	0.26	-	0.28	0.17	0.64	0.31					
		Male	0.64	0.82	0.01	0.006	0.009	0.25	14 0.87 - 0.80 0.26 0.63 58 0.04 - 0.38 0.38 0.77 53 0.19 - 0.59 0.59 0.82 14 0.53 - 0.03 0.35 0.94 06 0.92 - 0.78 0.22 0.25 03 0.86 - 0.64 0.48 0.02 028 0.26 - 0.28 0.17 0.64 028 0.26 - 0.28 0.17 0.64 028 0.26 - 0.28 0.17 0.64 0.28 0.16 0.41 0.85 0.16 0.41 88 0.61 - 0.23 0.48 0.81 53 0.19 - 0.59 0.59 0.82 14 0.53 - 0.03 0.35 0.94 71 0.44 - 0.87 0.94 0.82 <td>0.41</td> <td>0.11</td>	0.41	0.11			
Granulocyte-macrophage0.16Respiratory burst index*6.32Lysosomal presence1.84Disconstruction connectivity2.27	1.84×10^{-2}	Female	0.40	0.08	0.009	0.15	0.32	0.88	0.61	-	0.46	0.73	0.85	0.94
Lysosoniai presence	1.64 X 10	Male	0.79	0.32	0.84	0.94	0.32	0.40	0.92	-	0.23	0.48	October Mesocosme I 0.63 0 0.77 0 0.82 0 0.94 0 0.25 0 0.02 0 0.64 0 0.41 0 0.85 0 0.81 0 0.94 0 0.82 0 0.94 0	0.89
		Female	0.55	0.81	0.96	0.86	0.77	0.53	0.19	-	0.59	0.59	0.82	0.30
Phagocytosis capacity*	3.37 x 10 ⁻⁶	Male	0.21	0.93	0.89	0.07	0.01	0.14	0.53	-	0.03	0.35	0.94	0.37
****		Female	0.30	0.38	0.44	0.67	0.55	0.71	0.44	-	0.87	0.94	0.82	0.30
Phagocytosis efficiency*	7.64 x 10 ⁻⁸	Male	0.58	0.76	0.73	0.08	0.26	0.64	0.28	-	0.46	0.49	0.06	0.38

Table III: Values correspond to means \pm standard deviations (with n = 15-20) for each innate immunomarker tested. Statistical differences between *laboratory* and *mesocosm* or *field* data were tested by a Dunnett test (parametric data, $p \le 0.05$) or by a Kruskal-Wallis test followed by a Dunn test (non-parametric data, $p \le 0.05$). With *, ** and *** indicates significant difference at $p \le 0.05$, $p \le 0.01$ and $p \le 0.001$ levels respectively.

	DECEMBER			FEBRUARY		APRIL			JUNE			AUGUST			OCTOBER			
	Laboratory	Mesocosme	Field	Laboratory	Mesocosme	Field	Laboratory	Mesocosme	Field	Laboratory	Mesocosme	Field	Laboratory	Mesocosme	Field	Laboratory	Mesocosme	Field
Apoptotic cells	3.84 ± 2.21	3.73 ± 1.19	2.65 ± 1.15	4.79 ± 2.37	5.01 ± 1.73	4.94 ± 3.16	4.02 ± 1.39	5.61 ± 2.02	8.58*** ± 5.87	9.15 ± 6.36	6.75 ± 5.15	-	5.27 ± 2.50	4.63 ± 1.84	4.91 ± 2.51	5.22 ± 3.25	5.64 ± 2.66	4.68 ± 3.73
Necrotic cells	1.77	1.67	1.25**	2.39	1.88	1.20***	2.36	1.99	2.41	3.68	7.31***	-	1.86	1.60	0.79***	2.08	2.61**	1.75
(%)	± 0.54	± 0.67	± 0.40	± 1.81	± 0.54	± 0.60	± 0.92	± 0.42	± 1.06	± 8.87	± 2.62		± 1.05	± 0.51	± 0.67	± 1.90	± 0.94	± 0.88
Granulocyte-macrophage (%)	25.90 ± 10.32	25.30 ± 6.65	32.80* ± 9.66	27.20 ± 10.00	19.40 ± 4.93	39.30* ± 8.17	18.50 ± 6.21	17.70 ± 8.47	27.20*** ± 5.78	8.19 ± 13.26	34.00*** ± 6.33	-	28.60 ± 9.33	32.20 ± 9.83	32.40 ± 9.82	26.20 ± 7.40	28.40 ± 6.21	41.20*** ± 10.37
Respiratory burst index	2.11	3.27	1.99	1.27	1.33	1.57***	1.09	0.66	1.90***	3.62	1.36***	-	1.98	1.80	1.34***	1.27	1.07	2.13***
(experimental unit)	± 1.09	± 4.67	± 0.85	± 0.73	±0.71	± 0.74	± 0.32	± 0.31	± 1.08	± 3.71	± 0.77		± 0.91	± 0.43	± 0.36	± 0.77	± 0.26	± 1.06
Lysosomal presence	2.25	1.99**	2.30	2.31	2.56***	2.64***	2.40	2.67*	3.30***	2.16	2.18	-	2.17	2.15	1.52***	2.03	2.17	1.81*
(Mean Fluorescence Intensity)	± 0.27	± 0.27	± 0.42	± 0.28	± 0.26	± 0.42	± 0.29	± 0.21	± 0.36	± 0.26	± 0.17		± 0.34	± 0.24	± 0.23	± 0.35	±0.24	± 0.29
Phagocytosis capacity	55.90	65.90 ***	55.10	63.40	59.10	52.40***	62.30	63.60	62.40	65.00	60.10***	-	64.90	66.20	40.90***	69.50	67.70	64.70***
(%)	± 4.09	± 4.21	± 3.18	± 13.11	± 2.84	± 4.56	± 6.78	± 4.52	± 4.27	± 5.64	± 6.25		± 4.00	± 4.29	± 2.31	± 4.54	± 4.28	± 4.37
Phagocytosis efficiency	21.70	24.10	25.20	45.70	28.00***	18.70***	42.40	35.80*	23.90***	26.40	19.90***	-	12.80	14.30	18.10***	32.30	32.20	23.00 ***
(%)	± 11.25	± 3.81	± 3.06	± 16.78	± 2.60	± 3.62	± 6.95	± 5.08	± 4.60	± 5.13	± 5.17		± 2.44	± 2.28	± 1.71	± 6.65	±7.67	± 3.41

Table IV: Effects of sampling period, experimental level (*laboratory*, *mesocosm* or *field* data), sex and fish size (and interactions) on innate immune parameters of the three-spined stickleback, *Gasterosteus aculeatus*. Results were obtained by performing a type II ANCOVA ($p \le 0.05$). With *, ** and *** indicates statistically significant effect at $p \le 0.05$, $p \le 0.01$ and $p \le 0.001$ levels respectively in F-tests. R² were calculated on the model including only the significant explicative variables and interactions. Q²LOO corresponds to the squared cross-validated correlation coefficient obtained using a leave-on-out procedure. Biomarkers with * were log transformed before performing statistical tests.

	Explicative variables										Q ² LOO
	Period	Experimental level	Sex Size	Period:experimental level	Period:sex	Experimental level:sex	Period: size	Experimental level:size	Experimental level:Period:size	_	
Apoptotic cells*	***			***			***	*	***	0.33	0,24
Necrotic cells*	***	***	**	***	**					0.43	0,43
Granulocyte-macrophage	***	***	**	***	*		***	*		0.52	0,46
Respiratory burst index*	***	***	*	***						0.29	0,24
Lysosomal presence	***	**		***						0.53	0,51
Phagocytosis capacity*	***	***	***	***			***			0.63	0,60
Phagocytosis efficiency*	***	***	***	***			***	*	***	0.80	0,77

Table V: Proportion of total variability A) explained by period and its interactions (binary and ternary) with other confounding factors B) explained by the period:experimental level interaction and its interactions (ternary) with other confounding factors, C) explained by experimental level and its interactions (binary and ternary) with other confounding factors, for each immunomarker, with the data collected in mesocosms, with and without June, and for the data collected in field, compared to data collected in laboratory conditions. Biomarkers with * were log transformed before performing statistical tests

		Mesocosm			Mesocosm withou	ıt June	Field					
	Period (%)	Period: experimental level (%)	Experimental level (%)	Period (%)	Period: experimental level (%)	Experimental level (%)	Period (%)	Period: experimental level (%)	Experimental level (%)			
Apoptotic cells*	33.4	7.38	7.71	24.0	8.42	9.58	28.3	11.4	13.3			
Necrotic cells*	32.1	19.2	22.2	18.4	5.57	6.80	31.0	13.3	26.2			
Granulocyte-macrophage	45.1	18.4	21.8	33.9	7.80	11.0	27.1	3.09	19.1			
Respiratory burst index*	32.5	6.64	8.20	39.6	4.73	5.19	30.8	11.5	14.8			
Lysosomal presence	26.5	5.36	6.89	26.9	5.81	7.64	58.1	26.6	27.9			
Phagocytosis capacity*	40.0	15.3	16.3	41.2	13.8	16.2	45.6	23.8	40.8			
Phagocytosis efficiency*	79.5	5.54	6.32	81.3	4.89	5.21	77.8	18.8	23.1			

Laboratory reference ranges

Tested in field/mesocosm sampling

