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## Article

# Retrospective Modeling of NO<sub>2</sub> and PM<sub>10</sub> Concentrations over the Lyon Metropolitan Area (France), 1990–2010—Performance Evaluation, Exposure Assessment and Correlation between Pollutants

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**Abstract:** Numerous cancers develop years after subjects have been exposed to chemical compounds. Thus, environmental epidemiological studies need to accurately reconstruct exposures over long periods. To estimate exposure to NO<sub>2</sub> and PM<sub>10</sub> concentrations, we modelled ground-level air concentrations, at very fine temporal (1 h) and spatial (10 m) resolutions, over a large European metropolitan area and at subject's address of a French national cohort, for five different years (1990, 1995, 2000, 2005 and 2010). Model performances were assessed by comparing the annual modelled concentration against monitoring station measurements. As input data, we used background concentrations from a large-scale dispersion model. The relevance of our approach was assessed by comparing results in 2010, with a modelling using monitoring values as background data. The comparison with measurement data showed good performance of the model for the majority of the period, with a performance declined in 1990. Concentrations at the subject's residence decreased by 45% for PM<sub>10</sub> and 38% for NO<sub>2</sub>. The proportion of subjects exposed above the WHO recommendations declined from 100% to 50% for PM<sub>10</sub> and from 79% to 16% for NO<sub>2</sub>. The results of this study would provide a reference for future models to assess chronic exposures to PM<sub>10</sub> and NO<sub>2</sub> on a larger scale.

**Keywords:** air pollution; long-term exposure; exposure assessment; particulate matter; nitrogen dioxide

## 1. Introduction

Air pollution is a major public health concern and has been consistently linked to a range of adverse health effects. According to the Global Burden of Disease Study, air pollution was responsible worldwide for 19% of cardiovascular deaths, 24% of ischemic heart disease deaths and 21% of stroke deaths in 2015 [1]. In Europe, despite decreasing pollutant concentration levels over the last decades, the annual excess mortality rate from ambient air pollution is considered to be close to 790,000 (95% CI 645,000–934,000) [2].

In an urban context, air pollution is a complex mixture of multiple pollutants, including several agents classified by the International Agency for Research on Cancer (IARC) as carcinogenic or probably carcinogenic to humans such as particulate matter (PM) or diesel exhaust [3,4]. The two main pollutants in urban and peri-urban environments are PM and nitrogen dioxide (NO<sub>2</sub>). For both, there is evidence of effects on human health associated with exposure to concentration levels that are lower than current legislation thresholds [5–9]. These observations have raised questions about the contribution of long-term exposure to low-level PM and NO<sub>2</sub> concentrations and the occurrence of a variety of pathologies. Long-term exposure to PM has been associated, with the risk of lung cancer, with sufficient evidence for a causal link and bladder cancer [4]. Epidemiological studies have also suggested a positive association with breast cancer [10]. Concerning long-term exposure to NO<sub>2</sub>, previous studies have reported positive associations with breast [11–14] and colorectal cancer [15].

Given the long latency of cancer development, the reliable characterization of long-term exposure to these pollutants raises important challenges, since it requires the estimate of pollutant concentrations over several decades and a fine spatial resolution over the whole study period. Assessment of exposure in European cities prior to the year 2000, has to face two main problems: the lack of direct concentration measurements and the lack of reliable estimates of pollutant emissions by different source typologies (traffic, domestic heating and industry). The lack of concentration measurements severely limits the use of statistical models, such as Land Use Regression Models, whose calibration relies on the availability of a relatively large number of measurement stations [16]. The high uncertainty in the emission scenarios is instead reflected on considerable uncertainties in the concentration estimates provided by dispersion models. The accuracy of the output of such models is further affected by the difficulties in accurately estimating background concentration levels. Indeed, these are typically estimated by sub-urban or rural measurement stations, yet rarely available before 2000 or as output of regional dispersion models [17] whose estimates suffer from the aforementioned uncertainties in the emission scenarios.

As a consequence, the longer the study period, the more the lack of knowledge about historical air pollution levels may lead to misclassification of study subjects, resulting in uncertainty and bias in the risk estimates [18,19].

In order to address these features, this study presented the results obtained with an urban dispersion model (SIRANE) for the assessment of NO<sub>2</sub> and PM<sub>10</sub> concentrations in the Lyon Metropolitan Area (LMA) between 1990 and 2010. Historical pollutant concentration fields at a fine temporal (1 h) and spatial (10 m) resolution were reconstructed for five different years: 1990, 1995, 2000, 2005 and 2010. The aims of the study were: (i) to evaluate the accuracy of the model outputs (assessed by comparison with available field measurements) using different methods to estimate background concentrations; and (ii) to assess the exposure of study subjects of the French national E3N cohort living in the LMA during the 1990–2010 period in the framework of a case-control study on the effects of long-term air pollution exposure on breast cancer risk [20]. The E3N study (Etude Epidémiologique auprès de femmes de la Mutuelle Générale de l'Éducation Nationale) is an ongoing prospective cohort involving 98,995 French female volunteers, enrolled in 1990 and born between 1925 and 1950 [21].

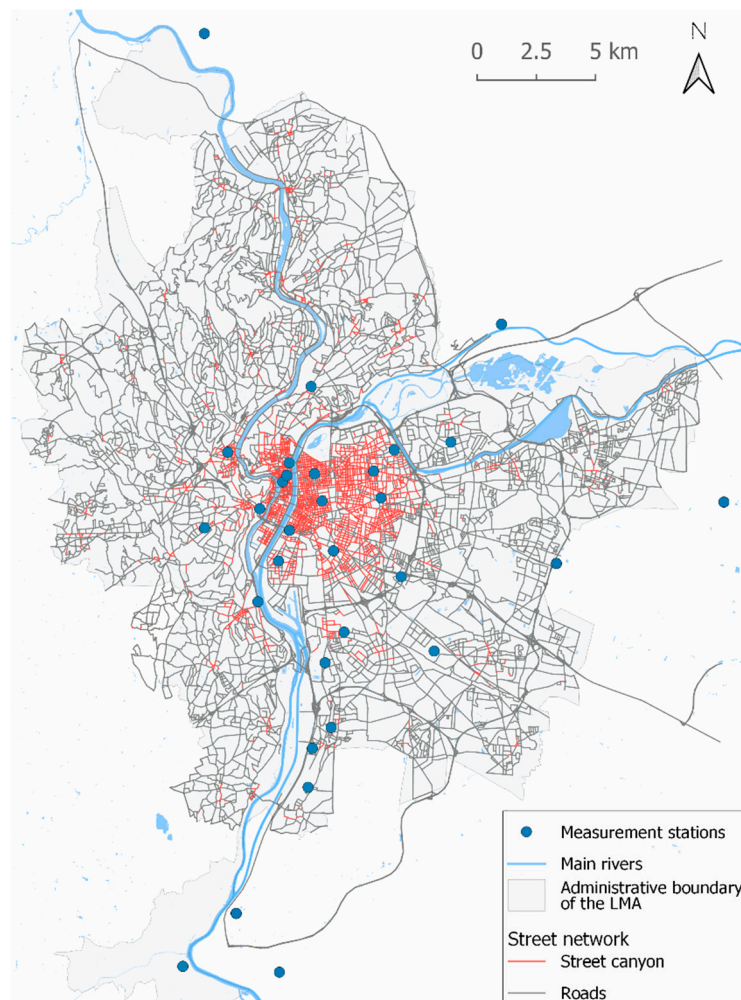
## 2. Materials and Methods

### 2.1. Study Area, Period and Population

The present study was conducted in the LMA, a 30 km × 34 km urban area (Figure 1), which has the second largest population in France (1.4 million people). The choice of the LMA in this present study was principally based on the availability of measurements of NO<sub>2</sub> and PM<sub>10</sub> concentrations in ambient air, from 1990 onwards, which was rare in the French context.

The domain includes a network (Figure 1) of 24,815 streets comprising 4985 streets classified as canyons (i.e., with a ratio width/height ≤ 3), which cover approximately 15%

of the total length of the street network. In 2010, 15 permanent monitoring stations of  $\text{NO}_2$  and 13 of  $\text{PM}_{10}$  (Figure 1) were located in the area. The number progressively reduces to only 5 stations, going back in 1990, for both pollutants.



**Figure 1.** The Lyon metropolitan area (LMA). The figure shows the locations of the measurement stations and street network (street canyons in purple lines).

For the present study, we selected 476 subjects from the E3N cohort study who lived in the LMA in 1990. The residential history was collected through self-administered questionnaires sent to the women in the cohort every 2 to 3 years. For the purpose of the study, we considered the residential addresses that remain unchanged over the whole study period (1990–2010) as done in a previous study evaluating model performance for dioxin and cadmium exposure assessment [22].

## 2.2. Atmospheric Dispersion Modeling

### 2.2.1. The SIRANE Model

The numerical simulations were performed with SIRANE, an operational model conceived to simulate pollutant dispersion at the local urban scale, assuming steady meteorological conditions over hourly time steps [23]. For that purpose, SIRANE represents the urban canopy as a network of connected streets whose relative pollutant exchange is modelled adopting ad hoc parameterizations, notably considering three exchange processes [24]: the advection of pollutant along the street axis [25], the dispersion at the intersections [26] and the ventilation through the turbulent exchange at the roof level [27].

Input data of the SIRANE model include the urban geometry, the meteorological data, the industrial, traffic and surface emissions (see Section 2.2.2) and the background concentrations (see Section 2.2.3). The only chemical reactions taken into account concern the cycle NO-NO<sub>2</sub>-O<sub>3</sub>, computed assuming a photo-stationary equilibrium [28]. To date, SIRANE has been extensively validated, both against wind tunnel experiments [29,30] or field measurements over the LMA [31,32] and other European cities [33].

All the simulations were carried out with the same street network (Figure 1), therefore ignoring potential modifications in the urban geometry and vehicle circulation that occurred over the study period. The meteorological conditions were reconstructed (with an hourly time step) from data registered at the Météo-France station at the Saint-Exupéry airport. Further details on the model are provided by Soulhac et al. [23,31,32].

### 2.2.2. Emissions

The emissions associated with traffic were represented as line sources. Emissions from road tunnel stacks and large facilities were represented as punctual elevated sources. Other miscellaneous emissions, such as domestic heating or small industries, were represented by a regular 1 × 1 km emission grid (See Supplementary Materials 1).

Emission data of NO<sub>2</sub>, NO and PM<sub>10</sub> were provided by Atmo Auvergne-Rhône-Alpes, the local authority for air quality as 1 × 1 km grid, in 2000, 2005 and 2010. Emissions were classified based on the activity sectors according to the Selected Nomenclature for Air Pollution (SNAP). To estimate the missing emissions for 1990 and 1995, we used a 3 × 3 km emissions grid, provided by the French National Institute for Industrial Environment and Risks (INERIS), from 1990 to 2000. We first established the emissions trends from 2000 to 1990 of each SNAP sub-category and for each cell in the INERIS grid and then applied these trends to the corresponding cells in the 1 × 1 km grid to estimate emission in 1995 and 1990.

### 2.2.3. Background Concentrations

SIRANE takes into account the contribution of hourly background concentrations, that is, due to emissions located outside the studied domain. The background concentration was considered as uniformly distributed over the study area and added to the concentrations estimated by SIRANE before applying the chemistry module (NO<sub>2</sub>-O<sub>3</sub>-NO). It can be estimated by means of concentrations measured at monitoring stations placed at the border of the domain and at distance from traffic axes [34–36] or by running a dispersion model over a larger (regional) domain [17,37,38].

As for the year 2010, measurement data were available for the background station as well as data from a large-scale dispersion model, the present study was able to compare the two approaches (Section 3.1). The background monitoring station (Saint-Exupéry) was located outside the domain at about 20 km from the Lyon city center (Figure 1) with available data on NO<sub>2</sub>, NO, O<sub>3</sub> and PM<sub>10</sub> for 2010. The large-scale dispersion simulations were performed by the INERIS with a Chemistry Transport Model, CHIMERE [39]. Hourly concentrations of NO<sub>2</sub>, NO, O<sub>3</sub> and PM<sub>10</sub> were simulated over France for years 1990, 1995, 2000, 2005 and 2010, with a spatial resolution of 0.125° × 0.6°.

It is worth noting that, given that no background monitoring station was available for all the studied years and to keep a consistent approach for all the studied years, the SIRANE simulation, described in Sections 3.2 and 3.3, used CHIMERE outputs as background concentrations.

## 2.3. Assessment Approach

### 2.3.1. Concentration Measurements

The annual concentrations modelled in 2010 using two different background data sets (measurement vs CHIMERE) were compared to 15 and 13 measurement stations for NO<sub>2</sub> and PM<sub>10</sub>, respectively.

Simulated annual averaged concentrations for the years 1990, 1995, 2000, 2005 and 2010, using CHIMERE estimates as background data, were compared to annual values from monitoring stations (Figure 1). Over the study period, the number of active measurement stations decreased going back in time: from 20 stations in 2010 to 5 in 1990 for NO<sub>2</sub> and from 13 to 5 stations for PM<sub>10</sub>. The distribution of monitoring stations in 1990 and 2010 for each pollutant are presented in Supplementary Material 2.

### 2.3.2. Subjects Exposure

To assess the exposure in the study population, annual average concentrations were computed at the residential addresses of the 476 E3N subjects, for five reference years (i.e., 1990, 1995, 2000, 2005 and 2010).

Firstly, we studied the evolution of NO<sub>2</sub> and PM<sub>10</sub> concentrations at the subject's address throughout the study period with the WHO recommendations for annual ambient air concentrations as a reference [40].

Secondly, we tested the correlation of exposures to a given pollutant between each year. This provided information about the reliability of data from a single year used to evaluate exposures over the entire period, as performed in many epidemiological studies [41–43].

Finally, knowing that some epidemiological studies, due to lack of data, used the exposure data of other pollutants as a proxy of exposure for their evaluation, we studied the capacity of NO<sub>2</sub> to predict PM<sub>10</sub> exposures over the study period and vice versa.

For these last two steps, the continuous and group rankings of the subjects were compared for each pollutant and each year. We used the Pearson correlation coefficient for continuous ranking and Cohen's Kappa for group ranking [44], two common statistical indicators used in epidemiological studies.

## 3. Results and Discussion

### 3.1. Influence of Background Concentration Data on 2010 Simulations

The present study assessed the performances of the SIRANE model, for 2010, using two different background concentrations as input data. The first simulation using measurements collected at the Saint-Exupery monitoring station and the second simulation using concentration estimated by the regional scale dispersion model CHIMERE at the Saint-Exupery monitoring station location.

The comparisons between annual mean concentrations measured and simulated for NO<sub>2</sub> and PM<sub>10</sub> are presented in Tables 1 and 2. Concerning the quality of the estimates at the monitoring stations for NO<sub>2</sub> (Table 1), both approaches showed relatively similar deviations (<20%) and correlation coefficients ( $R^2 > 0.80$ ). Concerning the performance of the modelling for PM<sub>10</sub> concentrations (Table 2), the SIRANE simulation using measured concentrations as background concentrations led to a better deviation. On the other hand, the correlation coefficients were similar to that observed for NO<sub>2</sub>.

**Table 1.** Comparison of NO<sub>2</sub> annual mean concentrations modelled with SIRANE ( $C_p$ ), using two set of background data and concentrations measured at the monitoring stations ( $C_m$ ).

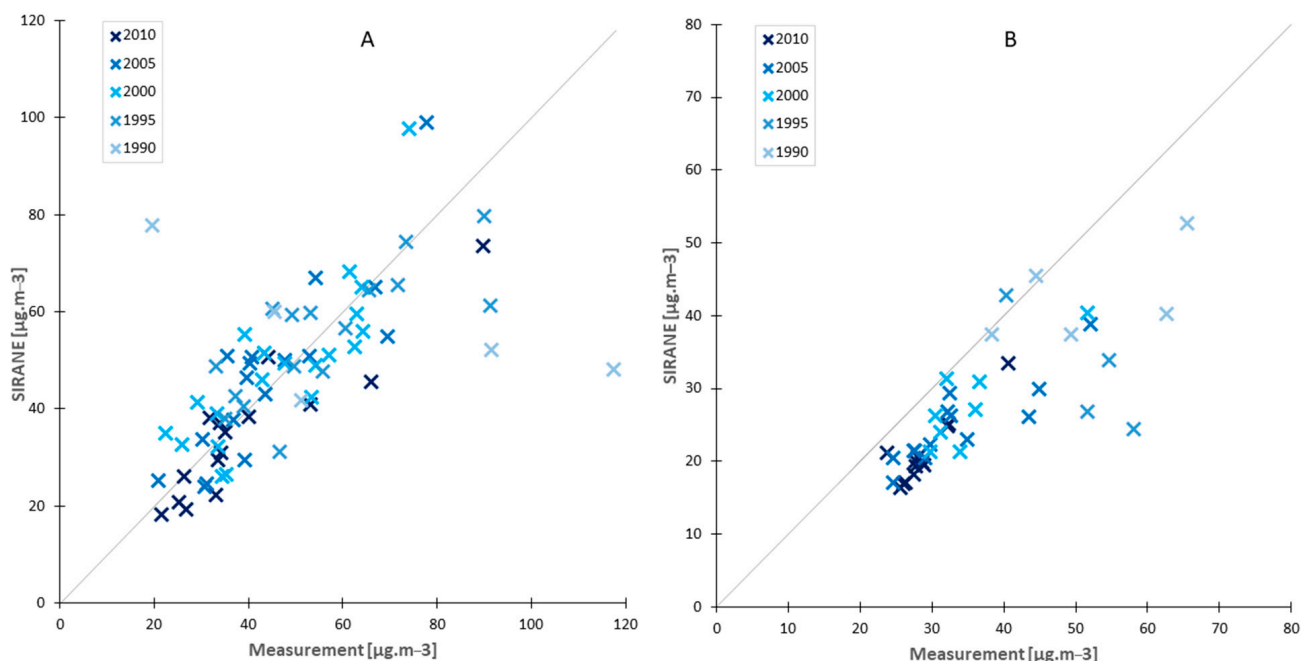
Background Data from the St Exupery Location	Number of Stations	$\overline{C_p}$ (SD)	$\overline{C_m}$ (SD)	$\frac{ \overline{C_p} - \overline{C_m} }{\overline{C_m}}$	$R^2$
Measured concentrations	15	33.5 (14.5)	39.7 (18)	18.3%	0.82
Simulated concentrations		35.2 (14.5)		15.7%	0.82

**Table 2.** Comparison of PM10 annual mean concentrations modelled with SIRANE ( $C_p$ ), using two set of background data and concentrations measured at the monitoring stations ( $C_m$ ).

Background Data from the St Exupery Location	Number of Stations	$\overline{C_p}$ (SD)	$\overline{C_m}$ (SD)	$\overline{\left(\frac{ C_p - C_m }{C_m}\right)}$	R <sup>2</sup>
Measured concentrations	13	30.1 (4.6)	28.9 (4.2)	4.8%	0.85
Simulated concentrations		21.0 (4.6)		27.7%	0.84

3.2. Comparison of Simulated Concentrations with Historical Annual Measurements

The comparison between modelled and measured NO<sub>2</sub> and PM<sub>10</sub> annual mean concentrations at the monitoring stations is presented in Figure 2.



**Figure 2.** Comparison between modelled and measured NO<sub>2</sub> (A) and PM<sub>10</sub> (B) annual mean concentration at the monitoring stations.

Concerning NO<sub>2</sub>, we observed good modelling performances except for the year 1990. With the exception of this year, the deviations from measurements were relatively constant (between 15% and 19%, Table 3). On the other hand, going back in time, we observed a decrease in the correlation between measured and modelled data with coefficients (R<sup>2</sup>) passing from 0.82 in 2010 to 0.59 in 1995.

**Table 3.** Comparison of NO<sub>2</sub> annual mean concentrations modelled with SIRANE ( $C_p$ ) and concentrations measured at the monitoring stations ( $C_m$ ).

Years	Number of Stations	$\overline{C_p}$ (SD)	$\overline{C_m}$ (SD)	$\overline{\left(\frac{ C_p - C_m }{C_m}\right)}$	R <sup>2</sup>
1990	5	56.0 (13.9)	65.0 (39.0)	90.4%	0.41
1995	16	55.0 (13.3)	56.0 (18.2)	15.8%	0.59
2000	20	48.9 (16.6)	47.1 (15.1)	19.2%	0.67
2005	17	47.2 (18.8)	44.6 (15.3)	17.3%	0.75
2010	15	35.2 (14.5)	39.7 (18)	15.7%	0.82

Concerning PM<sub>10</sub>, the modelled concentrations tended to be systematically underestimated compared to the measured ones. Except for the year 1995, the deviation from the measured concentrations varies around 25% (Table 4). Yet, the correlation between modelled and measured concentrations remain around 0.75 before falling sharply in 1990.

**Table 4.** Comparison of PM<sub>10</sub> annual mean concentrations modelled with SIRANE ( $C_p$ ) and concentrations measured at the monitoring stations ( $C_m$ ).

Years	Number of Stations	$\overline{C_p}$ (SD)	$\overline{C_m}$ (SD)	$\frac{ \overline{C_p} - \overline{C_m} }{\overline{C_m}}$	R <sup>2</sup>
1990	5	42.6 (6.5)	52.1 (11.7)	16.8%	0.33
1995	4	32.0 (8.3)	51.2 (7.67)	37.5%	0.75
2000	8	27.9 (6.4)	35.2 (7.1)	21.0%	0.71
2005	13	24.9 (5.6)	33.5 (8.4)	24.9%	0.76
2010	13	21.0 (4.6)	28.9 (4.2)	27.7%	0.84

It is worth noting that, for both pollutants, the worst performances were systematically observed, for earlier years, when measuring stations were less, that is, in 1990 for NO<sub>2</sub> and 1990 and 1995 for PM<sub>10</sub>. Results obtained for these 3 years should therefore be taken with caution.

Concerning the systematic underestimate of PM<sub>10</sub> concentrations, Nguyen et al. [45] pointed out the high contribution of the background to the PM<sub>10</sub> concentrations (over 50%). By comparing the PM<sub>10</sub> annual mean concentrations modelled with CHIMERE and measured at the Saint-Exupéry background monitoring station (available for 2005 and 2010, only), we observed that CHIMERE systematically underestimated measured PM<sub>10</sub> background concentrations (Table 5). This may therefore explain the systematic underestimation of the PM<sub>10</sub> annual mean concentrations.

**Table 5.** Comparison of PM<sub>10</sub> annual mean concentration modelled with CHIMERE ( $C_p$ ) and measured ( $C_m$ ) at the Saint-Exupéry background monitoring station for 2005 and 2010.

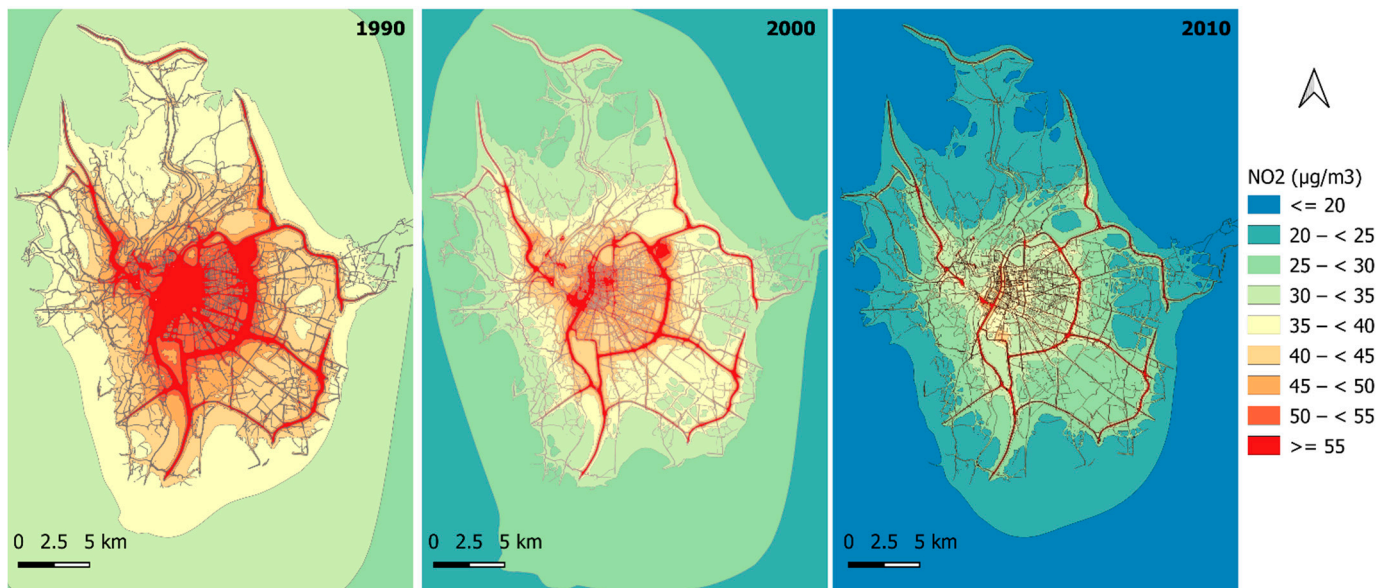
Years	$\frac{C_p - C_m}{C_m}$
2005	−0.28
2010	−0.37

### 3.3. Historical Trend in Annual PM<sub>10</sub> and NO<sub>2</sub> Concentrations

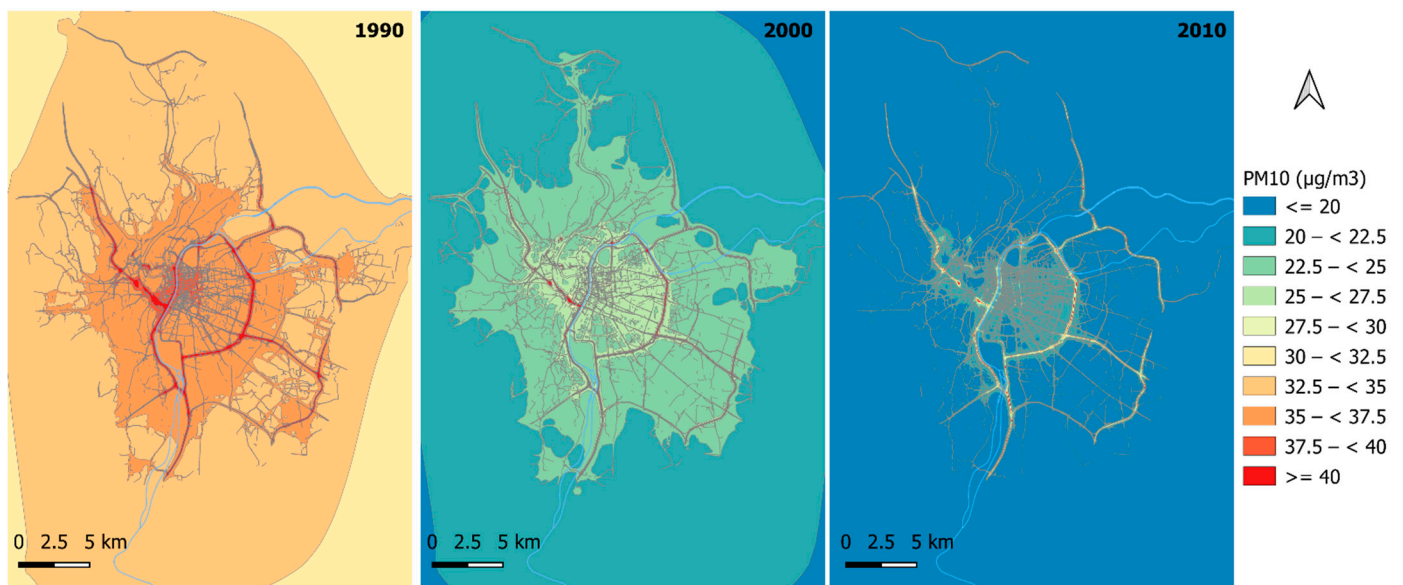
#### 3.3.1. Over the Study Domain

Figures 3 and 4 illustrate the evolution of the annual mean NO<sub>2</sub> and PM<sub>10</sub> concentrations simulated with the SIRANE model in the LMA from 1990 to 2010. These figures show a global reduction of the annual mean concentrations until 2010.

Concerning NO<sub>2</sub> concentrations (Figure 3), in 1990, all the city was exposed to annual concentrations exceeding the WHO recommendation (40 µg/m<sup>3</sup>) (Figure 3). Hotspots exceeding 90 µg/m<sup>3</sup> in annual concentrations were observed near major traffic roads. In comparison, in 2010, the areas where concentrations were above the WHO guideline were limited to the vicinity of major roads and in a few districts of the city center. In terms of hotspots, concentrations around 60 µg/m<sup>3</sup> were observed close to major roads.



**Figure 3.** Evolution over time of annual mean  $\text{NO}_2$  concentration ( $\mu\text{g}/\text{m}^3$ ), with a  $10 \times 10$  m resolution, in 1990, 2000 and 2010 in the LMA.



**Figure 4.** Evolution over time of annual mean  $\text{PM}_{10}$  concentration ( $\mu\text{g}/\text{m}^3$ ), with a  $10 \times 10$  m resolution, in 1990, 2000 and 2010 in the LMA.

For  $\text{PM}_{10}$ , in 1990, the entire study area was characterized by ground-level concentrations above the WHO recommendation ( $20 \mu\text{g}/\text{m}^3$ ). In district near the city center, the concentrations systematically exceed  $40 \mu\text{g}/\text{m}^3$ . Twenty years later, the situation greatly changed and most of the study area was characterized by the concentrations below the WHO guideline. Areas with concentrations slightly above this limit were found in the city center and there were a few places where concentrations were twice this limit (i.e., very close to major roads and tunnel exits).

These observations were confirmed by the study of annual simulated average concentrations, which have fallen in 20 years by 42% and 48% for  $\text{NO}_2$  and  $\text{PM}_{10}$ , respectively (Table 6). From 1990 to 2010, the percentage of the studied area with annual mean concentrations exceeding the WHO limit values decreased from 26.5% to 2.2% for  $\text{NO}_2$  and from 100% to 7.7% for  $\text{PM}_{10}$ . As regards the most exposed parts of the territory (99th percentiles

of annuals concentrations), the drop in concentrations were even more pronounced: from 74  $\mu\text{g}/\text{m}^3$  to 42  $\mu\text{g}/\text{m}^3$  for  $\text{NO}_2$  and from 54  $\mu\text{g}/\text{m}^3$  to 35  $\mu\text{g}/\text{m}^3$  for  $\text{PM}_{10}$ . All these observations witnessed a relevant improvement of the air quality over the study period.

**Table 6.** Annual mean concentration with standard deviation and percentage of the LMA with annual mean concentration exceeding the limit value mentioned by the WHO (40  $\mu\text{g}/\text{m}^3$  for  $\text{NO}_2$  and 20  $\mu\text{g}/\text{m}^3$  for  $\text{PM}_{10}$ ).

	$\text{NO}_2$		$\text{PM}_{10}$	
	Mean Annual Concentration ( $\mu\text{g}/\text{m}^3$ ) (SD)	Proportion of Study Area above WHO Recommendation	Mean Annual Concentration ( $\mu\text{g}/\text{m}^3$ ) (SD)	Proportion of Study Area above WHO Recommendation
1990	37.8 (11.1)	26.5%	33.8 (2.3)	100%
1995	36.0 (11.5)	22.9%	26.9 (2.5)	100%
2000	30.7 (9.6)	11.7%	21.9 (2.8)	95.1%
2005	28.2 (9.7)	9.0%	20.8 (2.6)	50.9%
2010	21.8 (7.1)	2.2%	17.5 (2.0)	7.7%

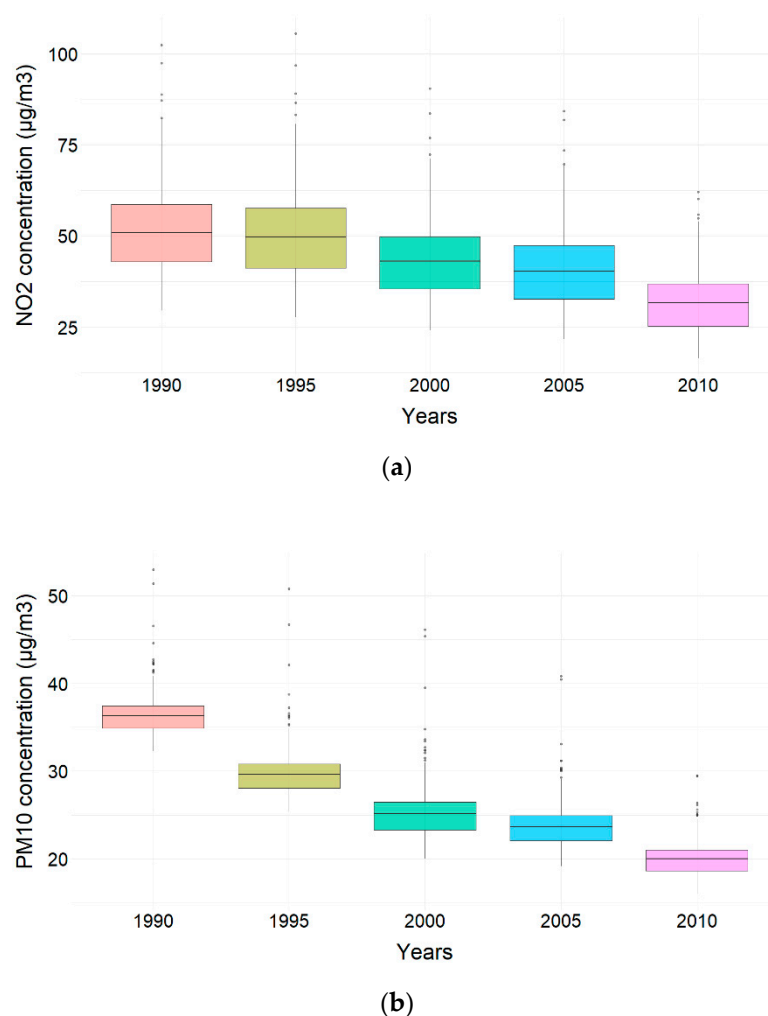
We noted that average  $\text{NO}_2$  and  $\text{PM}_{10}$  concentrations observed over the study area are similar to evolutions reported in long-term studies on air pollution. Sally Liu et al. [46] modelled the  $\text{NO}_2$  concentrations in Switzerland with a 200 m resolution and estimated exposure in 1993 and 2003 in the range of 7–76  $\mu\text{g}/\text{m}^3$  and 5–68  $\mu\text{g}/\text{m}^3$ , respectively. Yanosky et al. [47] estimated the  $\text{PM}_{10}$  exposure across USA from 1988 to 2007 and observed mean concentration ranging from 25.1  $\mu\text{g}/\text{m}^3$  for the 1988–1998 period to 21.5  $\mu\text{g}/\text{m}^3$  for the 1999–2007 period.

### 3.3.2. $\text{PM}_{10}$ and $\text{NO}_2$ Concentrations at the Study Subjects' Address of Residence

The annual median modelled  $\text{NO}_2$  concentrations at the subjects' address of residence decrease by 38%, that is, from 51.0  $\mu\text{g}/\text{m}^3$  in 1990 to 31.6  $\mu\text{g}/\text{m}^3$  in 2010. Whereas in 1990, 79% of the study subjects were exposed to  $\text{NO}_2$  levels above the WHO limit, only 16% were exposed above this value in 2010 (Table 7). Similar to the decrease of modelled concentrations over the entire LMA, the decrease in concentrations at subjects' addresses occurs in two stages, first between 1995 and 2000, then in 2005 and 2010. We noted that a few subjects were exposed to extreme values (above 80  $\mu\text{g}/\text{m}^3$ , twice the WHO recommendation) in 1990 and 1995 and that these subjects remain exposed above 50  $\mu\text{g}/\text{m}^3$  until 2010 (Figure 5). All of these hotspots had in common to be very close to roads (but not necessarily to major traffic roads).

**Table 7.** Annual median concentration with standard deviation at the subject location and proportion of subject addresses exposed above the World Health Organization (WHO) guidelines (40  $\mu\text{g}/\text{m}^3$  for  $\text{NO}_2$  and 20  $\mu\text{g}/\text{m}^3$  for  $\text{PM}_{10}$ ).

	$\text{NO}_2$		$\text{PM}_{10}$	
	Median (Min–Max) Concentrations at Subject Addresses	Proportion of Subject Exposed above WHO Recommendation	Median (Min–Max) Concentrations at Subject Addresses	Proportion of Subject Exposed above WHO Recommendation
1990	51.0 (29.5–102.4)	79%	36.3 (32.2–53.0)	100%
1995	49.6 (27.8–105.6)	76%	29.6 (25.3–50.8)	100%
2000	43.1 (24.1–90.5)	59%	25.1 (20.0–46.0)	100%
2005	40.2 (21.7–84.2)	51%	23.7 (19.2–40.8)	93%
2010	31.6 (16.5–62.1)	16%	20.0 (16.0–29.5)	50%



**Figure 5.** Boxplot of annual concentrations, NO<sub>2</sub> (a) and PM<sub>10</sub> (b), simulated at the subjects' location over time.

Over the 20-year period, median PM<sub>10</sub> concentrations at the subjects' locations decreased by 45%, from 36.3 µg/m<sup>3</sup> to 20.0 µg/m<sup>3</sup>. It was observed that the variability of PM<sub>10</sub> exposure in subjects is significantly lower than that of NO<sub>2</sub> exposure (Figure 5).

In 1990, all the E3N subjects were exposed to values above the WHO limit for PM<sub>10</sub> whereas in 2010, half of them are exposed to concentrations exceeding this value (Table 7). The main reduction occurred between 1990 and 2000 with an 11 µg/m<sup>3</sup> drop in annual concentration. Some high concentration value (>40 µg/m<sup>3</sup>, 2-time the WHO recommendation) was therefore observed for a couple of E3N subjects until 2005 (Figure 5). It is worth noting that the addresses with extreme PM<sub>10</sub> concentrations were the same as for NO<sub>2</sub>. This suggests that despite the PM<sub>10</sub> concentrations in the area was reduced and traffic was not the major source of PM<sub>10</sub> emissions, the highest exposures were still located close to roads.

The results presented here unveil a general trend of air pollution levels that were similar to those presented in other previous studies, even though these were generally focusing on larger domains and with a lower spatial and temporal resolution. Notably, some parallels have been drawn with studies conducted on long-term exposure to NO<sub>2</sub> and PM<sub>10</sub> in European cohorts. In a study on long-term residential exposure to PM<sub>10</sub> and NO<sub>2</sub> in a Danish cohort, with a resolution of less than 1 km, Hvidtfeldt et al. [48] found median (5th–95th) concentrations of 25.0 (17.9–39.5) for NO<sub>2</sub> and 25.1 (23.7–30.3) for PM<sub>10</sub>, over the 1993–2015 period. In the French context, Bentayeb et al. [49] conducted a study on residential exposures in the Gazel cohort between 1989 and 2008. For NO<sub>2</sub>,

they observed a 40% decrease in concentrations with a median (min-max) value going from 25 (2–94) to 15 (1–71). For PM<sub>10</sub>, median (min-max) concentrations decreased from 26 (6–44) µg/m<sup>3</sup> to 17 (4–29) µg/m<sup>3</sup>, that is, a 27% drop. Despite the lower resolution used in the Bentayeb study (2 km), the results and trends were comparable to the results presented in this study (Table 7). The study carried out on the ELFE (Étude Longitudinale Française depuis l’Enfance) cohort in France in 2010 and 2011 by Riviere et al. [50] was instead based on concentration fields, the resolution of which was similar to that adopted here. Their estimates of the average exposure of subjects (21 µg/m<sup>3</sup> for PM<sub>10</sub> and 24 µg/m<sup>3</sup> for NO<sub>2</sub>) were fairly close to ours in 2010 for PM<sub>10</sub> (20.0 µg/m<sup>3</sup>), less for NO<sub>2</sub> (32 µg/m<sup>3</sup>).

### 3.3.3. Correlation of Subject’s Exposure over Years

We observed that subjects’ exposures to NO<sub>2</sub> for the different years of the study period are highly correlated with each other, with a Pearson’s coefficient close to 1 (Figure 6). The results were similar for PM<sub>10</sub> with a Pearson’s coefficient always equal to or greater than 0.95. This implies that any year of the study period was a very good indicator of the subject’s exposure for any other years. In addition, these results also indicated that, in a densely populated urban area, NO<sub>2</sub> is a pertinent predictor of PM<sub>10</sub> exposure and vice versa, regardless of the year of the study period (minimum coefficient value: 0.93). Among the five years considered here, the best results were however obtained for the most recent year (2010) for both pollutants.

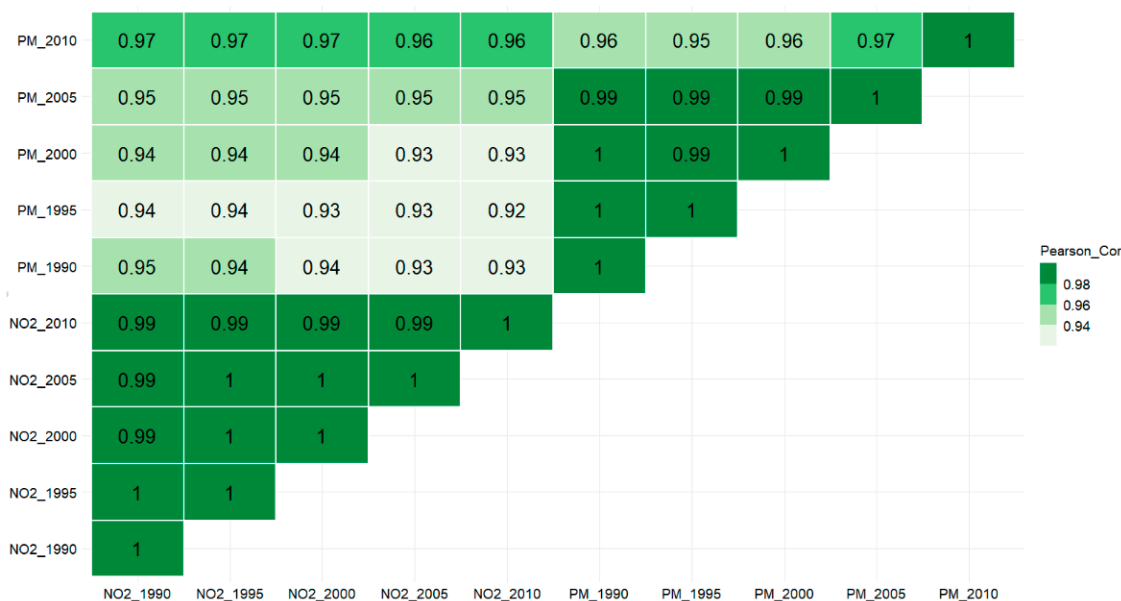


Figure 6. Pearson correlation coefficient between subjects’ exposure to PM10 and subject’s exposure to NO<sub>2</sub> (using Pearson correlation coefficient).

A previous study of exposure to air pollution in France from 1989 to 2007 [49] estimated Pearson correlation coefficients between air pollutant concentrations at a 2 km resolution. As observed in our study, their results also showed a strongly similar spatial distribution of NO<sub>2</sub> concentrations over the different years studied (correlation coefficients above 0.97). On the other hand, the correlation observed between PM<sub>10</sub> concentrations was less strong than in our study (from 0.63 to 0.87). Similarly, the correlations between PM<sub>10</sub> and NO<sub>2</sub> concentrations were less marked than in our study (around 0.74).

The analyses of the group classification agreement (Figure 7) also showed good results. For NO<sub>2</sub>, the agreement (kappa coefficients) between the years varied from 0.81 to 0.95, while for PM<sub>10</sub>, the values varied from 0.79 to 0.94 with only two pairs of years below 0.8 (1990 vs. 2010 and 2010 vs. 2000). A Kappa coefficient between 0.6 and 0.8

indicates a “strong” agreement between the two classifications, while a value above 0.8 indicates an “almost perfect” agreement. The agreement between the PM<sub>10</sub> and NO<sub>2</sub> group classifications varies from 0.70 to 0.81. For both pollutants individually, the weakest agreements were observed for the year 2000, whereas for the NO<sub>2</sub> vs PM<sub>10</sub> comparisons, the strongest agreement was observed in 2000.

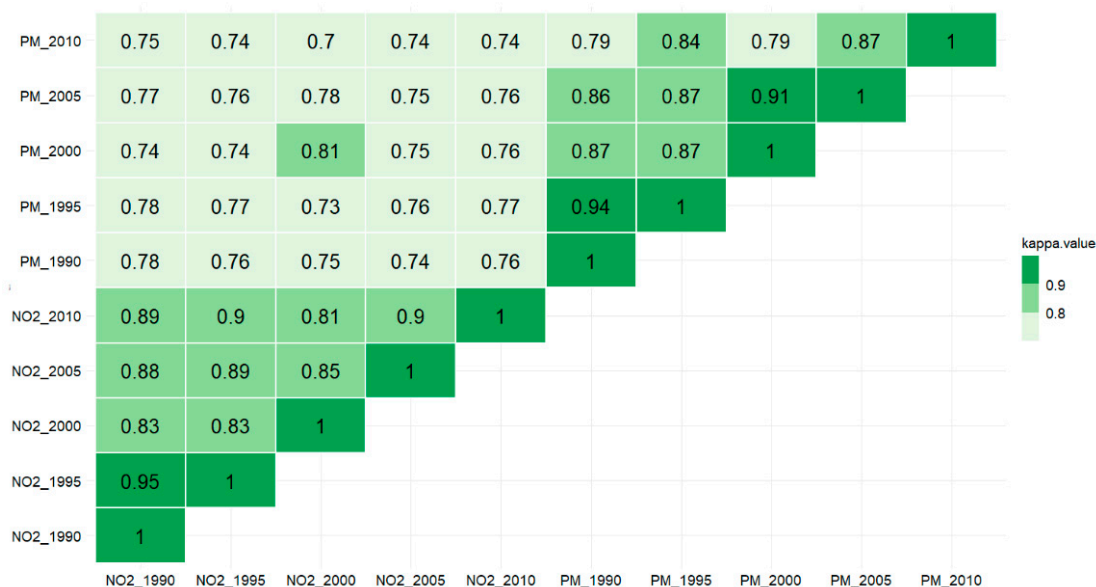


Figure 7. Agreement on the classification of subjects into five groups according to their exposure (using the Cohen’s kappa).

#### 4. Conclusions

In the framework of a case-control study on the effects of long-term air pollution exposure on breast cancer risk [20], we modelled ground-level air concentrations of PM<sub>10</sub> and NO<sub>2</sub> over a large European metropolitan area and over the 20 year period at a very fine resolution.

To our knowledge, this is the first study assessing the spatiotemporal variability of PM<sub>10</sub> and NO<sub>2</sub> concentrations over two decades in a large urban area, with a fine spatial resolution (10 × 10 m), using an urban dispersion model. Modelling results were compared to the available field measurements over the whole study period and were subsequently used to assess exposure at individual residence locations.

A preliminary analysis concerned the role of different typologies to be used as background concentration in the numerical simulations. Notably, for the year 2010, only the background concentration was provided by field measurements and by the output of a regional dispersion model (CHIMERE). Our local-scale modelling results show similar performance when focusing on NO<sub>2</sub> concentrations. Larger discrepancies can be instead observed for PM<sub>10</sub> annual concentrations, which were shown to be underestimated when using CHIMERE data (−27%) as the background value. Despite this, our results show that, for 2010, both background data provide good performances for the SIRANE model.

The rest of the simulations over the twenty year study period were then performed CHIMERE data as background, since no measurements were available as background data. The comparison with field measurements showed that our modelling results were able to correctly predict annual concentrations over the majority of the study period. For NO<sub>2</sub>, except the year 1990, the mean deviation remained under 20%. For 1990, the large deviation observed could be related to the uncertainties inherent in the emission data before 2000 also to the large reduction in the number of available measuring stations (from 16 in 2000 to 5 in 1990). For PM<sub>10</sub>, the median deviation varied between 16.8% and 37.5%. A larger gap was observed for the year 1995. As for NO<sub>2</sub>, this could be explained by the uncertainties

in the emission data and by the very low number of measuring stations (4 in 1995). We observed a high correlation that tended to decrease with the years from 0.82 to 0.59 for NO<sub>2</sub> and from 0.84 to 0.71 for PM<sub>10</sub>.

We noted a systematic underestimate of PM<sub>10</sub> concentrations all along the study period, which, as already observed in 2010, directly reflects the underestimation of background concentration as obtained by the CHIMERE output. We can therefore assert that our concern of producing results with a homogeneous methodology over the whole period, using modelled background values instead of measured values (unavailable after 2005), led to a lower model's performance for PM<sub>10</sub>.

As expected, over the period 1990–2010, we observed a decrease in modelled concentrations (−48% in PM<sub>10</sub> concentrations; −42% in NO<sub>2</sub> concentrations). This decrease could be explained by the reduction in emissions due to technological improvements in emission control and increasingly strict regulations. This was a welcome development: the part of the territory exposed above the WHO recommendations has fallen from 26% to 2% for NO<sub>2</sub> and from 100% to 7.7% for PM<sub>10</sub>. Nevertheless, in 2010 there were still a few hotspots in the LMA with concentrations above the WHO.

The study of the exposure of the members of the E3N cohort at their residential addresses makes it possible to refine these results. Although the proportion of the territory exposed to high concentrations has decreased, the most populated areas still are the areas most exposed to air pollution. Thus, even though the exposure at the subject's residential addresses has fallen sharply (−30% NO<sub>2</sub>; 45% PM<sub>10</sub>), still 16% and 50% of addresses were exposed to concentrations exceeding the WHO guidelines for NO<sub>2</sub> and PM<sub>10</sub> in 2010.

To reconstruct exposures related to a specific pollutant, it is useful to know its degree of correlation with other pollutants perhaps more measured. It is interesting to note that, in our study, these were the same subjects' addresses that were exposed to extreme concentrations of PM<sub>10</sub> and NO<sub>2</sub> over the years. Indeed, we observed a high degree of correlation of subjects' exposures to PM<sub>10</sub> and NO<sub>2</sub> with Pearson coefficients ranging from 0.92 to 0.97. Likewise, the comparison of the exposure group rankings showed a "strong" agreement. Thus, in this study case, the NO<sub>2</sub> data were good predictors of PM<sub>10</sub> exposure and vice versa.

One of the main challenges faced by epidemiological studies is to accurately characterize chronic exposures. This is particularly important when these studies concern pathologies that develop with a latency period in relation to exposure (as for cancers). When exposure data are lacking, it is common practice to use exposure data from a single year and to make the assumption that these exposure data are representative of the whole study period [51]. In our case study, for both pollutants, the classification of subjects according to their exposure changed slightly over the years (Pearson coefficient > 0.94). The study of the exposure group rankings of the subjects showed a degree of agreement that can be described as "almost-perfect." This indicates that, in this study, any year would be a good indicator of exposure for the study period for a given pollutant.

As a final step, it is worth mentioning the main limitations to our work. Firstly, the lack of background pollution measurement data over the whole period forced us to use modelled values as background values, which worsen the modelling results. This degradation is likely to be due to the uncertainties in providing reliable PM emissions data in the large-scale model. Secondly, the lack of local data on emissions for the decade 1990–2000 led us to use emission data with a lower spatial resolution (3 × 3 km). Thirdly, the number of available measurement data was very low in 1990 for both pollutants and in 1995 for PM<sub>10</sub>, which limited the analysis of the quality of the simulations before 2000.

At this stage, the results obtained in this study cannot be generalized and require to be confirmed by similar studies over other urban areas (with similar spatial and time resolutions). In particular, it would be interesting to see whether the high degree of correlation between exposures to PM<sub>10</sub> and NO<sub>2</sub> is actually influenced by the relatively small size of the territory studied and by the resolution of the model. This should be

investigated in future researches by comparing multiple models in the same area and studying the implications on the misclassification of individuals' exposure.

The information provided by our numerical results on the historical exposure of the E3N cohort subjects will be used to assess the accuracy of future national size models at the local scale and will improve the results of epidemiological studies.

**Supplementary Materials:** The following are available online at <https://www.mdpi.com/2073-4433/12/2/239/s1>, Supplementary materials 1 include Figure S1: Spatial distribution of the traffic sources for NO<sub>2</sub> and PM<sub>10</sub> in 2010, Figure S2: Spatial distribution of the punctual sources for NO<sub>2</sub> and PM<sub>10</sub> in 2010 and Figure S3: Spatial distribution of the surface distributed sources for NO<sub>2</sub> and PM<sub>10</sub> in 2010. Supplementary materials 2 include Figure S4: Spatial distribution of NO<sub>2</sub> monitoring stations in 1990 and 2010 and Figure S5: Spatial distribution of PM<sub>10</sub> monitoring stations in 1990 and 2010.

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